Photocatalytic Oxidation of Pollutant Dyes in Wastewater by TiO₂ and ZnO nano-materials – A Mini-review

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Abstract

In recent years, photocatalytico oxidation processes with ultra violet (UV) radiation and semiconductor photocatalyst like titanium dioxide (TiO₂) and zinc oxide (ZnO) have gained immense research interest as an effective wastewater purification method because of its efficacy in decomposing and mineralising the hazardous organic pollutants as well as the opportunity of utilizing the solar UV and visible spectrum. A comprehensive review of the UV assisted heterogeneous photocatalytic oxidation process is conducted with a border perspective into the photocatalytic reaction mechanism involved, different form of TiO₂ and ZnO catalysts, effective application towards waste water treatment by immobilized on surface or as suspension. Photocatalytic decomposition with ZnO and TiO₂ catalysts is mainly applied for treating organic contaminants dyes in wastewater because of their ability to attain fully mineralization of the organic contaminants under mild reaction conditions such as ambient pressure and temperature. One hundred and eighty seven published papers are reviewed and summarized here with the center of attention being on the photocatalytic oxidation of organic dyes present in waste water effluent. Most recently, photocatalytic degradation of oraganic molecules using TiO_2 and ZnO have been studied extensively due to the degradation of persistent organic pollutants, PAH (Polycyclic Aromatic Hydrocarbons) dyes and other organic chemicals which are known to be endocrine disruptors. Treatment of wastewater in a titanium and zinc dioxide suspended reactor has been widely used due to its simplicity and enhanced photodegradation efficiency.

Keywords : ZnO nanofiber, TiO_2 nanoaprticles, photocatalysis, wastewater, oxidation.

1. Introduction

The significance of the wastewater treatment, management and its disposal gradually increases in the modern times and it becomes a major concern for public health scientific interest. All existing protocols for treatment of wastewater are categorized as physical, chemical and biological processes¹. The sequential and concurrent use of those processes combinedly tends to create a greater efficient method in removing the pollutant aspects in liquid residues. Restrictions in terms of execution, efficiency, and price are a factor, however, biological processes, as an example, have been extensively used and show potential towards dairy and agricultural wastewater treatment². The chemical process deals with the photocatalysts like TiO₂, ZnO etc. mediated degradation of the industrials waste waters³. These processes have limitations which can potentially affect degradation efficiency through control pH range, rapid organic-load variations, and also the effluent's physicochemical behavior⁴.

The use of a heterogeneous photocatalyst is a conventional method for water purification which includes reduction and oxidation reactions from adsorbed wastewater, oxygen molecules and hydroxyl anions, or other organic molecules⁵⁻⁸. Uses of semiconductors like TiO₂, ZnO etc. in photocatalysis employ semiconductors in suspension⁹. However, this method could be a more expensive when it is scaled-up because particle-recovery of the photocatalyst particles is a difficult task and leads to an amplification in process costs. A feasible alternative is the preparation of photocatalyst layers in different substances or uses the catalytic support without hampering the photocatalyst activity. Many efforts have been made in which few studies have demonstrated continuous flow reactors with fixed-bed photocatalyst¹⁰. Integral to this study was an assessment of the efficiency of heterogeneous photocatalyst processes for dairy and agricultural wastewater treatment with immobilized TiO2 and ZnO to reduce organic pollutant load¹¹. The immobilization was carried

out by the application of a coating containing the catalysts¹²/photocatalyst¹³ or by suing the porous carbon or silica as a supporter for the catalysts to prevent agglomeration of catalyst during catalytic/photocatalytic reactions in solution. Moreover, solar emission was used as UV source¹³⁻¹⁵.

Recent review is focused on the most important photocatalysts titanium dioxide and zinc oxide and their photocatalytic activity towards wastewater treatment. Recently, Xiaobo and Samuel reviewed the broad applications of titanium dioxide as a photocatalysts¹⁶. These applications were comprised of photodecomposition of various industrial pollutants, killing tumor cell and killing bacteria in cancer treatments¹⁷. Semiconductor catalysts TiO₂ and ZnO have been widely used to mineralize harmful organic pollutants in wastewater into less damaging inorganic nontoxic compounds like CO₂, HCl and water¹⁸. Several studies have been carried out for decolorization of industrial wastewater by using photocatalysis and bacteria treatment¹⁹⁻²¹. The elimination of color from wastewaters is more necessary than the removal of other colorless organic compounds¹⁹. Because of aesthetic and environmental concerns the decolorization of effluent from textile dyeing and finishing industry has given most importance^{22,23}.

TiO₂/ZnO photocatalysis, in the presence of UV irradiation can disintegrate the pollutant dyes into non-toxic simple compounds like CO₂, HCl and water²⁴. Nanosized TiO₂ and ZnO photocatalysts in the form of nanorods, nanospheres, thin porous films, nanofibers and nanowires have been utilized in various applications, including photocatalysis because of their high activity, low cost and environmental safety²⁵⁻²⁷. Interestingly, very high surface to volume ratio of nanostructures make them efficient for photocatalysis and other application. In recent studies^{13,24}, authors have reported that zinc oxide and titanium dioxide have excellent photocatalytic properties and both catalysts are designated to be capable substrates for photodegradation of dyes water pollutants as they exhibit the acceptable activity in the range of ultraviolet radiation.

Various studies have focused on treatment of industrial wastewater using different treatment methods; however, most of these treatments have intricacy in realistic uses^{5,9,26,28,29}. In recent year, investigation on different systems have been carried out, such as, advanced oxidation processes (AOP)³⁰, ozonation^{31,32}, sonolysis³³, gamma–radiolysis³⁴, electro-coagulation^{35,36}, H₂O₂/UV³⁷, photocatalysis²⁴, photo-Fenton³⁸, biological and combined anaerobic–photocatalytic treatment^{21,29,39}. The aim of the present investigate is to investigate photocatalytic oxidation process for the decomposition wastewater using TiO₂ and ZnO as photocatalysts irradiated with artificial ultraviolet radiations.

Photocatalytic oxidation processes³⁰, which involve the generation of highly reactive hydroxyl radical (HO), have emerged as a promising water and wastewater treatment technology for the degradation or mineralization of a wide range of organic contaminants^{30,40,41}. The photoactivated reactions are characterized by the free radical mechanism initiated by the interaction of photons of a proper energy level with the catalyst (TiO₂, ZnO semiconductor catalysts)^{42,43}. The efficiency of a photocatalytic system is also affected by the form of TiO₂ and ZnO nanoparticle catalysts used as immobilized on surface or as colloidal suspension⁴⁴⁻⁴⁶. The photocatalysis reaction effective for the degradation of various organic impurities in waste water; however, its practical application as slurry type suspensions is limited due to the difficulty in separating the nanocatalysts particles after the photocatalytic reaction⁴⁷⁻⁴⁹.

The present review aims to provide a comprehensive analysis on the mechanism of UV-TiO₂ and ZnO photocatalytic oxidation process, photocatalyst material, irradiation sources, effect of pH, temperature, dye concentration, catalyst mass and type of catalysts on photocatalysis and the application towards wastewater treatment.

2. Photocatalytic Oxidation Process

In a typical photocatalytic system, photo reaction or photoinduced molecular transformation occurs on the photocatalyst surface³⁷. The basic mechanism of photocatalytic reaction is governed by the generation of electron-hole pair into the semiconductor and its transportation to destination (i.e., reaction with organic pollutants). Frank and Bard⁵⁰ first examined the possibility of using a titanium dioxide-TiO₂ catalyst for the decomposition of cyanide, Ollis and co-workers⁵¹ extensively studied the potential application of photocatalysis for organic degradation⁵². Semiconductor photocatalysts generally absorb different colour light depending on their bandgap energy and used photocatalysts because of their interesting electronic as configurations, light absorption ability, charge carrier transport property, and excited-state lifetimes^{25,53-55}. The principal reaction mechanism of a semiconductor photocatalyst (Fig. 1) is described as follows. When a photocatalytic surface is exposed by a radiation of energy equal to or greater than the bandgap energy (bandgap

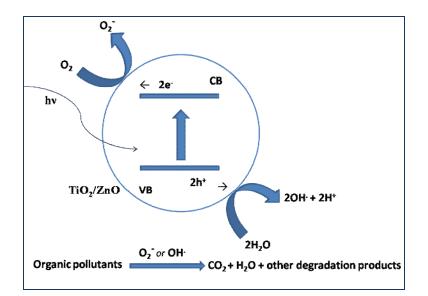


Fig. 1. Schematic of the photocatalytic mechanism for TiO₂/ZnO photocatalysts.

energies of common semiconductors are given in Table 1) of the semi-conductor photocatalyst material, it creates a positively charged hole in the valance band and negatively charged electron in the conduction band by exciting the electrons in the valance band to the conduction band⁵⁵.

Photocatalysts
$$(TiO_2/ZnO) + hv \rightarrow e^- + h^+$$
 (1)

$$h^{+} + H_2 O \rightarrow H^{+} + O H^{\bullet}$$
⁽²⁾

$$h^{+} + OH^{-} \rightarrow OH^{-}$$
(3)

$$e^{-} + O_2 \rightarrow O_2^{-} \tag{4}$$

$$2e^{-} + O_2 + 2H^+ \rightarrow H_2O_2 \tag{5}$$

$$e^{-} + H_2 O_2 \rightarrow OH^{-} + OH^{-}$$
(6)

Organic+•OH $+O_2 \rightarrow CO_2+H_2O+$ other degradation productions (7)

The conduction band electron reduces oxygen (into O_2) adsorbed to photocatalyst surface (TiO₂/ZnO) whereas the positively charged hole oxidizes either organic pollutants directly or indirectly by water to produce hydroxiyl free radicals (HO·). The photocatalytic reaction mechanism shown above (Figure 1) can be described by Eq. (1)-(7) as shown below^{25,56-58}; In the photocatalysis of the organic pollutants, the oxidation of pollutants and the reduction process of oxygen do not occurs concurrently, there is an accumulation of electrons in the conduction band of the photocatalyst, thus helping a recombination of negatively charged electron and positive hole. Therefore, efficient utilization of electrons is necessary to encourage photocatalytic oxidation process. Semiconductors can also be used as photocatalysts for environmental remediation. Here, the photon-induced electronhole pairs helps in redox reactions at the semiconductors catalyst surface and produces hydroxyl free-radicals (OH•) and superoxide ions (O₂). These generated species act as powerful oxidizer to disintegrate harmful organic pollutants in wastewater and convert them into CO_2 and H_2O . The most significant and fundamental

components for a proficient photocatalytic system are the photocatalyst, the light source the reactor configuration 25 .

Semiconductors	Bandgap energy (eV)
Diamond	5.4
Cu ₂ O	2.172
ZnS	3.6
TiO ₂	3.03
SnO ₂	3.54
ZnO	3.36
CdSe	1.7
WO ₃	2.76
Si	1.17
Fe ₂ O ₃	2.3
PbS	0.286
PbSe	0.165
ZrO ₂	3.87
Ge	0.744
CdS	2.42

Table 1. Bandgap energies for some common semiconductor materials at 0 $K^{54,59}$.

The photocatalytic oxidation reaction depends on generation and recombination of electrons and holes in the semiconductor photocatalyst. Adsorbed oxygen on photocatalyst surface acts as an

electron captor, which can control the recombination of photo induced electron-hole recombination. The nanosized catalysts have larger photocatalytic efficiency as compared to than normal photocatalyst materials^{60,61}. The reasons behind this are:

(i) **Quantum Size Effect :** One of the most direct effects happens when the particle size becomes less than a certain critical limit (*i.e.*, size reduces to the nanometer range), this give rises to quantum size effects due to the confinement of the electronic movement. This leads to the valence band and conduction band of the semiconductor change into discrete energy levels, depending on the size of the material structure, which means that the electric potential of valence band changes more positive; otherwise the conduction band electric potential changes more negative. Then the oxidation-reduction potential of the electrons and holes is increased, and therefore the oxidation activity of nanosized TiO₂ and ZnO photocatalyst is enhanced^{62,63}.

(ii) **Higher Specific Surface Area:** If there are more number of atoms resides on the surface, it enhances the adsorption capacity of the photocatalysts towards organic pollutants. The activity of photocatalysis is associated with the time exhausted by electrons and holes to get to the surface of the particles. When the particles are nano sized their diameter becomes very miniature, so it is then very straightforward for the charge carriers moving from the inside to the surface, and starting red-ox reaction. Higher the surface to volume ratio, smaller the particle diameter and the smaller time may be used up by charge carriers diffusing to the surface from inside. This can provide lesser probability of electron and hole recombination. Therefore, the superior photocatalytic activity can be achieved. Consequently, the nano-titanium and zinc oxides have greater photocatalytic efficiency than the common (bulk) titanium or zinc oxides⁶⁴⁻⁶⁶.

3. TiO2 and ZnO as Efficient Photocatalyst for Wastewater

3.1. TiO₂ as a Photocatalyst

Being a member of metal-oxide semiconductor photocatalysts family, there is a general agreement among researchers that TiO_2 is more superior because of its high phocatalytic activity, large chemical stability and robustness against photocorrosion, low price, and nontoxicity⁶⁷. Particularly, titanium dioxide is largely used for many applications such as air and water purification, as an agent for antifogging and self-cleaning surface, whereas, zinc oxide has not been used like TiO_2 for such applications^{13,68}. Wu has studied that TiO₂ has greater photocatalytic activity than ZnO and SnO_2^{69} . For the decomposition of phenol as target organic species, Okomoto studied that TiO₂ has greater photocatalytic activity as compared to cadmium sulfide (CdS)⁷⁰. Sakthivel proved that under similar study conditions, TiO₂ is a superior photocatalyst than WO₃, α -Fe₂O₃, SnO₂ and ZrO₂⁷¹. On the other hand, Augugliaro established that, TiO₂ has better photochemical stability than ZnO in aqueous solution; however, zinc-oxide is photochemically more active (although its surface to volume ratio was less)⁷². Significant amount of work has been carried out for investigating the photocatalytic activity of titanium dioxide based on its phase composition, size and crystal structure $^{73-75}$. Three different phases of TiO_2 exist: anatase, rutile and brookite⁷⁶. There are still very few reports on brookite while anatase and rutile phases of nanostructured TiO_2 are the most studied phases⁷⁷⁻⁷⁹. The position of oxygen ions on the exposed anatase TiO₂ particle surface possesses a triangular arrangement which allows significant absorption of organic molecules⁸⁰. Whereas, the orientation of titanium ions in the anatase TiO₂ creates an advantageous reaction condition with the absorbed organic pollutants⁸¹. Interestingly, these favorable structural arrangements of oxygen and titanium ions are not present in the rutile phase of TiO₂. Due to this reasons anatase phase of titania shows higher photocatalytic activity than rutile⁸². Although, anatase phase is reported to be the most photochemically active form of titaniumdioxide, but several researchers believed that a pure anatase phase has less photocatalytic performance⁸³. The rutile phase present in the anatase TiO₂ introduces some percentage of mesoporosity and a wider pore size distribution which probably be accountable for the enhanced photocatalytic activity for this phase⁸⁴. These reports have been suggested that anatase-rutile mixtures/composites would be the best combination to synergistically enhance photocatalytic efficiency⁸⁵⁻⁸⁷.

Various strategies have been taken to prepare commercial samples of TiO_2 by varying purity, particle size, and crystal structure and phase composition to enhance photocatalytic efficiency with Degussa P25 grade (generally, it is a mixture of 30% rutile and 70% anatase TiO_2 powder samples)^{88,89}. There is a general believe that commercial grade Degussa P-25 TiO₂ sample shows significant photocatalytic degradation efficiency towards organic dye degradation as compared to other commercial forms^{89,90}. The effect of particle size on the photocatalysis can be explained in terms of available surface area of TiO₂ for the reaction. Usually, for the smaller particle size, surface area would be larger and surface to volume ration increases which intern increases expected photocatalytic activity. This larger available surface area thus increases the number of active photocatalytic sites per unit area, as well as enhances the adsorbability of the organic pollutants on the TiO₂ photocatalyst surface⁹¹. However, several shortcomings of the conventional TiO₂ powder catalysts, such as difficulty in separation after synthesis and stirring throughout reaction, make powder TiO₂ less photocatalytic efficient and prevent form being used in the practical applications⁹². Hence, several studies have been focused on the synthesis and modification of TiO₂ nanostructured by making composite photocatalysts with ceramics like SiO₂, ZrO₂, Al₂O₃ and

TiO₂ coated polymer nanostructures to eliminate the previous difficulties⁹³⁻⁹⁵. Tryba *et al.* has also been reported that TiO₂-mounted activated carbon could increase the photocatalytic degradation of organic dyes⁹⁶. Interestingly, different form of carbons like porous carbon, carbonnanotube (CNT)⁹⁷, carbonnanofiber (CNF)⁹⁸, graphene⁹⁹, exfoliated graphite¹⁰⁰ and graphene oxide¹⁰¹ have been reinforced with TiO₂ to make TiO₂/carbon composites for superior photocatalytic activity.

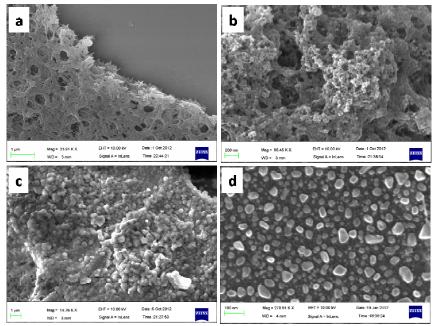


Fig. 2 : Degradation of RhB dye contaminated water with $TiO_2/porous$ carbon films. The porous structure results in improved catalytic support for the photocatalyst as discussed in [13] :Fe-SEM micrographs of the $TiO_2/polymer$ and $TiO_2/carbon$ porous films: (a) Porous polyacrylonitrile (PAN) polymer and TiO_2 composite film, (b)Higher magnification images of 'a' showing the presence of TiO_2 nanoparticles on the polymer films, (c) Porous carbon and TiO_2 composite films after pyrolyzation of $TiO_2/polymer$ film, and (d) Higher magnification images showing 'bead' like titanium dioxide particles on the porous carbon film.

Even though, in photocatalysis anatase phase of titania has more efficiency than rutile but, this phase is thermodynamically

metastable¹⁰² and it may undergoes irreversibly transformation from anatase to rutile phase at relatively high temperatures ~ 800° C and above¹⁰³. Moreover, studies have shown that rutile pahse of TiO₂ exhibit superior photocatalytic activity compared to pure anatase TiO₂ for the photodecomposition of *p*-coumaric acid¹⁰⁴, photo-oxidation of H₂O with Fe^{3+,105} and photodecomposition of H₂S¹⁰⁶. Besides that, rutile TiO₂ has better light scattering properties, chemical inertness and cost efficient which makes it a potentially candidate for several applications like dye sensitized solar cells, solar photocatalytic, electrochemical and catalysis¹⁰⁷⁻¹¹⁰.

In a recent study¹³, authors have developed an *in-situ* synthesis method for pure rutile TiO_2 nanoparticles impregnated macroporous carbon film and found efficient for the photocatalysis of Rhodamine B dye. This may be advantageous for the applications of TiO_2 when there is a requiring of pure rutile phase and the working temperature is higher.

3.2. ZnO as a Photocatalyst

Till to date, many types of semiconducting systems have been studied for photocatalysis including TiO₂, ZnO, ZrO₂, CdS, WO₃, and so on¹¹¹⁻¹¹⁶. Most of them have band gap in the UV (ultraviolet) region, *i.e.*, equal to or greater than 3.36 eV ($\lambda = 388$ nm)⁵⁹. Thus, these catalysts promote photocatalytic reactions under the illumination of UV radiation¹¹⁷. Surface area, surface defects and bangap of the metal oxide nanostructured catalysts play a significant role in the photocatalysis¹¹⁸. 1D- nanostructures like nanowires, nanorods, nanospheres and nanofibers are potential candidates for application to photocatalysis because they have offered a very high surface to volume ratio than nanoparticles^{62,119,120}.

Zinc oxide is an excellent wide band gap, natural n-type semiconducting material with binding energy (60 MeV), abundant in nature, nontoxic nature and environmental friendly photocatalyst¹²¹. It can absorb wider range of spectrum of radiation which also makes it more applicable for dye sensitized solar cells and solar photovoltaic applications¹²²⁻¹²⁴.

The nanosized ZnO oxides are most efficient functional materials for photocatalytic applications which are available at low cost, exhibit mild reaction conditions and high photochemical reactivity, while affording the use of sunlight¹²⁵. Amongst all metal oxide semiconductors, zinc oxide, most functional material find application for many fields including biosensors, light-emitting diodes, field-emission, and gas sensing¹²⁶.

The photocatalytic efficiency of ZnO nanomaterials is believed to be much better than normal photocatalysts¹²⁷. It is extensively used to treat wastewater, such as printing and dyeing wastes, dairy and food wastewater, drugs and pesticides wastewater, textile wastewater, papermaking wastewater, and so on¹²⁸⁻¹³².

There is growing interest for the photocatalytic degradation of different organic pollutants present in wastewaters comes from industrial and agricultural waste^{5,24,31,92,133}. The photocatalytic reactions happens in such treatment somehow identical to heterogeneous catalysis which includes the simultaneous adsorption of oxygen and organic reactant species present on the waste, followed by the oxidation on the ZnO photocatalyst surfaces^{6,10}. Zinc oxide is a photoactive semiconducting material oxide and able to activate itself by taking energy for the photocatalytic reaction from photons. Therefore ZnO photocatalyst must have to simultaneously adsorb reasonable amount of O₂ and reducing species (i.e., the organic pollutant molecules). A good photocatalyst like zinc oxide is considered to provide adsorption sites for the organic pollutants, which means that an open porous structure with high specific surface area is needed¹³⁴. Since the thermal energy is expensive, a very important issue of a good photocatalysis is that the involving reactions have to occur at ambient temperature.

Recently, it is proven that semiconducting photocatalyst mediated photocatalytic oxidative degradation of organic compounds is a doing well¹³⁵. This has been demonstrated as an alternative to conventional wastewater treatments for the removal of organic pollutants from various industrial wastes¹³⁶. The uses of ZnO nanostructures as a potential photocatalyst material to decompose the environmental pollutants have also been rigorously studied^{137,138}. Interestingly, to achieve better photocatalytic efficiency in many realistic applications, it is expected that ZnO photocatalyst should absorb not only ultraviolet radiation but also visible light¹³⁹. To absorb visible spectrum of solar radiation, band gap energy of ZnO photocatalysts must have to be decreased or divided into different sub-bandgaps, which can only be obtained either by incorporating transition metal ions or by doping nitrogen, sulphur atoms¹⁴⁰. The electronic and photocatalytic properties of zinc oxide solely depended on its surface morphology, size, crystal structure, aspect ratio, density of crystal and crystallographic orientation^{141,142}. ZnO has known as an effective photocatalyst for water detoxification, organic pollutant decomposition and other photolysis because it produces hydrogen peroxide (H_2O_2) more proficiently¹⁴³. Also, it has high mineralization and reaction rates, and also provides more of active sites with high specific surface photoactivity¹⁴⁴. So far, many attempts have been taken to create different morphology of nanosized ZnO photocatalysts in which hydrothermal synthesis¹⁴⁵, sol-gel technique¹⁴⁶, plasma-enhanced chemical vapor deposition (PECVD)¹⁴⁷, Rf-magnetron sputtering¹⁴⁸, electrospinning^{24,149} and so on¹⁵⁰⁻¹⁵². Recently, we have reported²⁴ a novel synthesis protocol for the fabrication of mesoporous ZnO nanofibers by employing electrospinnig technique. In that paper, we have demonstrated an ultraviolet radiation assisted photodecomposition of PAHs dyes-naphthalene and anthracene by using free standing mesoporous ZnO nanofibers mats.

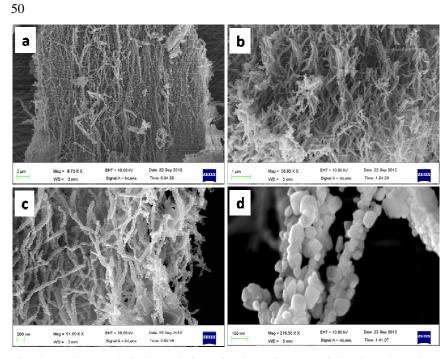


Fig. 3 : Photocatalytic degradation of polycyclic aromatic hydrocarbon dyesnaphthalene and anthracene in wastewater water with electrospun mesoporous ZnO nanofiber mats as discussed in [24]: Fe-SEM micrographs of the electrospun ZnO nanofibers: (a) free-standing fiber mats, (b) partially aligned nanofibers, (c) Higher magnification of 'b' and (d) a single mesoporous nanofiber.

4. Effect of Light Source on Photocatalytic Degradation

The percentage degradation of dyes in waste water improved with increasing exposed light intensity¹⁵³. Under the elevated intensity of light irradiation, the enhancement was significantly higher since the electron–hole formation is predominant at high irradiation intensity and, therefore, electron–hole recombination probability is insignificant¹⁵⁴. However, when irradiated light intensity becomes very poor, separation of electron–hole pair competes with recombination which consecutively decreases the formation of free radicals, thus, causing less result on the percentage degradation of the waste water¹⁵⁵⁻¹⁵⁷.

The greater part of photocatalytic degradation studies have been carried out between the wavelengths 320-380 nm⁵⁹ which is related to the bandgap energy of the TiO₂ and ZnO photocatalyst¹⁵⁸. Thus, the light source which produces the required radiation field can be performed by solar irradiation or by artificial lamps. In a typical photocatalytic reactor, radiation (wavelength ranging between 320-380 nm) is provided by fluorescent low-pressure and medium mercury lamps emitting low and high intensity UV light, respectively in the short, medium and long UV spectrum⁵⁹. The higher possibility of trapping of electronhole pairs with shorter wavelength excitation and direct photolysis were considered to be likely the reasons for a good photocatalytic system¹⁵⁹. Only \sim 5% of the incident solar radiation is used by TiO₂ and ZnO assisted photocatalytic decomposition of dyes^{59,160}. This drastically confines their practical application. Therefore, modification of TiO₂ and ZnO photocatalysts to improve light absorption capacity and photocatalytic activity under visible light irradiation is the topic of recent research¹⁶¹⁻¹⁶³.

5. Effect of pH on Photocatalytic Degradation:

The pH value of the aqueous solution is a key parameter for photocatalytic degradation of wastewater and dyes because it affects the adsorption of pollutants that happens at the surface of photocatalysts^{164,165}. The wastewater produces from textile, dairy and pharmaceutical industries generally have a wide range of pH values. In addition, the generation of hydroxyl radicals which is necessary for the photocatalytic reaction also depends on pH of the solution¹⁶⁶. Therefore, pH plays a significant role for both cases, in the chemical nature of wastes and generation of hydroxyl radicals^{1,167}. Hence, many attempts have been taken to investigate the effect of pH in the degradation of wastewater in the ultraviolet and solar irradiations¹⁶⁸⁻¹⁷⁰. Photocatalytic decomposition of dyes in waste water have been studied at pH values ranging from 3

(acidic) to 13 (alkaline) for all the three (anionic, cationic and neutral) dyes in wastewater¹.

It was observed that, at optimal concentration of dyes in both acidic and alkaline pH tends to lower the degradation efficiency of some azo dye wastes^{171,172}. It was studied that in presence of H₂O₂ nonbiodegradable cationic dyes some (Malachite green, Rhodamine B, and Methylene blue) degrades least effectively at both acidic and alkaline pH, but highly effective at neutral pH value of the solution¹⁷³. The inhibitory effect found to be more prominent in the high alkaline range (*i.e.*, pH 11-13)¹⁷¹. At high pH values the hydroxyl radicals are so quickly discarded that they do not have proper time to react with dyes in waste water. The pH influences the surface properties of TiO₂ and ZnO photocatalysts, dissociation of organic pollutants (or dyes in waste water) and formation of hydroxyl radicals²⁵.

The greater photodegradation efficiency at neutral pH can be explained by the point of zero charge (pzc) of the photocatalysts¹⁶⁹. The pzc value was found at pH 6.25^{170} and 8.9^{174} for the TiO₂ and ZnO respectively. The photocatalysts surfaces become negatively charged under alkaline conditions (pH > pzc), whereas it is positively charged in acid media (pH< pzc). While TiO₂ and ZnO exhibit an amphoteric nature with a zero charge in the pH range around their pzc value, the adsorption of dye pollutants is extremely good at that pH value, and consequently, higher photodegradation activity was predictable at a neutral pH value¹⁶⁹. Therefore, the result indicates that pH value has a significant effect on the adsorption properties at the photocatalyst surface and hence photodegradation rates are quite insignificant with extreme pH values¹⁷⁵.

The interpretation of effects of pH values on the activity of the photodegradation process is a very complicated task, because there are three possible reaction mechanisms behind the organic pollutant/dye degradation process, namely, oxidation and reduction by the positive hole and negatively charged electron, respectively

in the semiconductor photocatalysts and hydroxyl radical attack. The importance of each parameter strongly depends on the pH value and nature of the substrate. In the case of organic dyes and pollutants, it can be assumed that the main reaction is governed by the hydroxyl radical attack, which can be advantageous by the high hydroxyl radicals' concentration at around neutral pH values of the solution¹⁶⁹. Additionally, another justification for the pH effects can be connected with ionic specification of the organic dyes because the protonation or deprotonation of the dyes can drastically alter its adsorption characteristics and oxidation-reduction activity^{1,171}.

6. Effect of Temperature, Dye Concentration, Catalyst Loading and Type of Catalysts on Photocatalysis:

In a typical photocatalytic decomposition of wastewater dyes, the following operating parameters have also involved in the process.

6.1. Temperature : Abass *et al* studied¹⁹ that decolorization of real textile industrial wastewater with time increases with increasing temperature for all types of catalysts. Their results confirm that higher temperature is significantly helpful to decompose the pollutants in wastewater. This is probably due to the fact that the activation energy gets increased with increasing operating temperature¹⁶⁹.

6.2. Dye Concentration : The lowering in dye (pollutant in waste water) concentration reduces the time of decomposition of the wastewater. This performance could be described as: when the concentration of real industrial wastewater is very less than of the original concentration then the catalysts active sites probably be completely exposed by dye ions¹⁷⁶. The further increase in dye concentration may also be responsible for screening the exposed light so the light intensity will be reduced¹⁷⁷.

6.3. Photocatalyst Loading : The effect of photocatalytic degradation of wastewater dyes with different catalyst loading has been extensively explored^{18,171,178,179}. Konstantinou and his coworkers have shown that, for azo dye degradation by TiO₂ assisted photocatalysis, the initial degradation rates of azo dyes in aqueous solution were directly proportional to TiO₂ catalyst concentration 171 . Interestingly, studies for UV assisted photocatalytic decomposition of industrial wastewater using different photocatalyst (anatase or rutile or zinc oxide) loading shows that for an optimum photocatalysts mass, which was required for utmost decolorization efficiency does not dependent on exposure time, type of irradiation source and characteristics of pollutant¹⁸⁰. Further studies have also revealed that the decolorization efficiency increases with increase in catalysts loading and it get saturated at a certain value of loading and then it starts decreasing with further increasing the mass^{19,181}. The increasing photocatalytic activity with increase in photocatalysts's happens may be due to increasing accessibility of masses photocatalyst sites and the diminish of catalytic activity after the saturation region is correlated to rising of light scattering by the much excess of photocatalyst particles^{181,182}.

6.4. Type of Catalysts : Considering the impact of different prospective photocatalyst on their thermal and chemical activity, stability under different working environment, accessibility and ease of handiness in many physical forms, cost effectiveness, toxicity, and environmental friendly the common widely used photocatalyst is TiO_2^{31} . Amongst the three crystal structures, only rutile and anatase are established enough and can be used as a potential photocatalyst¹⁸³. Pillai and his co-worker has reported that photocatalytic properties of different phases of TiO_2 materials might be dissimilar significantly and rutile phases exhibiting the lowest photoactivity^{184,185}. Anatase phase of TiO_2 is superior photocatalyst than rutile while brookite phase has not studied much¹⁸⁶. In the case of Degussa P-25, a mixture of anatase and

rutile TiO_2 which prevents the recombination rate of photogenerated electrons and holes and for this reason shows more activity than anatase¹⁸⁷.

Further, Hussein and Abass¹⁹ studied the decomposition of real industrial wastewater on rutile, anatase, and zinc oxide at a constant temperature and time and has proven that the efficiency of these catalysts fell in the sequences as shown below:

 $ZnO > Degusa P-25 > TiO_2 (Anatase) > TiO_2 (Rutile)$

Interestingly, the quantity of ZnO photocatalyst required to achieve the optimal photocatalytic activity is nearly double than that for TiO₂ (anatase or rutile)^{19,182}.

7. Conclusions

Reviewing the recent representative publications, the function operating of various parameters on the photocatalytic decomposition of various ognaic dyes in wastewater explored in this review. TiO₂ and ZnO have been recommended to be efficient photocatalysts for the degradation and mineralisation of various toxic organic pollutants such as azo dyes in wastewater water. The investigations also suggest that the coexistence of photocatalyst and lights exposure is necessary for photocatalytic degradation of dyes. Various operating parameters such as nature light source, pH of the reaction medium, temperature, dye concentration, catalyst loading and type of catalysts have a considerable effect on degradation efficiency of dyes in wastewater. Optimization of the photodecomposition parameters is essential from the viewpoint of efficient design and the application of photocatalytic oxidation processes to guarantee sustainable wastewater purification process.

Although the UV energy gives improved efficiency in degradation of dyes, solar irradiation could appear as a substitute cost effective light source because of its abundance. Thin-film coating of photocatalyst may resolve the problems of leaching and separation. Make the naofibers of catalysts may further enhance the photocatalytic activity due to its high surface to volume ration which will facilitate the better adsorption of dyes. Metal and nonmetal doped nanostructured TiO_2 and ZnO have been reported to result in improved degradation rates due to their modified band gap energy for using visible and solar radiation. So, we need to focus for developing more reliable photocatalysts which can absorb visible and solar radiation or by both. In addition, further work is essential on the designing and understanding the working parameters for oxidation of pollutant dyes in wastewater.

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References

- Neppolian B, Choi HC, Sakthivel S, Arabindoo B, Murugesan V. (2002) Solar/UV-induced photocatalytic degradation of three commercial textile dyes. Journal of Hazardous Materials, 89(2-3), 303–17.
- Sin J-C, Lam S-M, Mohamed AR, Lee K-T. (2012) Degrading endocrine disrupting chemicals from wastewater by TiO₂ photocatalysis: A review. International Journal of Photoenergy, 185159.
- Nakata K, Fujishimaa A. (2012) TiO₂ photocatalysis: Design and applications. Journal of Photochemistry and Photobiology C: Photochemistry Reviews, 13(3), 169-89.
- Vilar VJP, Gomes AIE, Ramos VM, Maldonado MI, Boaventura RAR. (2009) Solar photocatalysis of a recalcitrant coloured effluent from a wastewater treatment plant Photochemical & Photobiological Sciences, 8(5), 691-8.
- Chong MN, Jin B, Chow CWK, Saint C. (2010) Recent developments in photocatalytic water treatment technology: A review. Water Research, 44(10), 2997-3027.
- Rizzo L, Meric S, Guida M, Kassinos D, Belgiorno V. (2009) Heterogenous photocatalytic degradation kinetics and detoxification of an urban wastewater treatment plant effluent contaminated with pharmaceuticals. Water Research, 43(16), 4070–8.

- 7. Abhang RM, Kumar D, Taralkar SV. (2011) Design of photocatalytic reactor for degradation of phenol in wastewater. International Journal of Chemical Engineering and Applications, 2(5), 337-41.
- 8. Ibhadon AO, Fitzpatrick P. (2013) Heterogeneous photocatalysis: Recent advances and applications. Catalysts, 3, 189-218.
- Ni M, Leung MKH, Leung DYC, Sumathy K. (2007) A review and recent developments in photocatalytic water-splitting using TiO₂ for hydrogen production. Renewable and Sustainable Energy Reviews, 11, 401–25.
- Samanamud GRL, Loures CCA, Souza AL, Salazar RFS, Oliveira IS, Silva MB, et al. (2012) Heterogeneous photocatalytic degradation of dairy wastewater using immobilized ZnO. ISRN Chemical Engineering, 275371, 1-8.
- Homem V, Santos L. (2011) Degradation and removal methods of antibiotics from aqueous matrices - A review. Journal of Environmental Management, 92(10), 2304-47.
- 12. Mondal K, Kumar J, Sharma A. (2013) Self-organized macroporous thin carbon films for supported metal catalysis. Colloids and Surfaces A: Physicochemical and Engineering Aspects, 427, 83–94.
- Mondal K, Kumar J, Sharma A. (2013) TiO₂ nanoparticles impregnated photocatalytic macroporous carbon films by spin coating. Nanomaterials and Energy.
- 14. Rizzo L, Fiorentino A, Anselmo A. (2013) Advanced treatment of urban wastewater by UV radiation: Effect on antibiotics and antibiotic-resistant E. coli strains. Chemo sphere, 92(2), 171–6.
- 15. Priya SS, Premalatha M, Anantharaman N. (2008) Solar photocatalytic treatment of phenolic wastewater-potential, challenges and opportunities. ARPN Journal of Engineering and Applied Sciences, 3(6), 36-41.
- Chen X, Mao SS. (2007) Titanium dioxide nanomaterials: Synthesis, properties, modifications, and applications. Chemical Reviews 107(7), 2891-959.
- Attia AJ, Kadhim SH, Hussen FH. (2008) Photocatalytic degradation of textile dyeing wastewater using titanium dioxide and zinc oxide. E-Journal of Chemistry, 5(2), 219-23.
- 18. Akpan UG, Hameed BH. (2009) Parameters affecting the photocatalytic degradation of dyes using TiO_2 -based photocatalysts: A review. Journal of Hazardous Materials 170(2-3), 520-9.

- 19. Hussen FH, Abass TA. (2010) Photocatalytic treatment of textile industrial wastewater. International Journal of Chemical Science, 8(3), 1353-64.
- Asad S, Amoozegar MA, Pourbabaee AA, Sarbolouki MN, Dastghei SMM. (2007) Decolorization of textile azo dyes by newly isolated halophilic and halotolerant bacteria. Bioresource Technology, 98(11), 2082-8
- Ndasi NP, Augustin M, Bosco TJ. (2011) Biodecolourisation of textile dyes by local microbial consortia isolated from dye polluted soils in ngaoundere. International Journal of Environmental Sciences, 1(7), 1403-19.
- 22. Elahee K. (2010) Heat recovery in the textile dyeing and finishing industry: lessons from developing economies Journal of Energy in Southern Africa, 21(3), 9-15.
- 23. M. RK, K.S. (2010) Advanced treatment of textiles yarn dying waste water towards resue using reverse osmosis membrane. International Journal on Applied Bioengineering, 4(1), 25-33.
- Singh P, Mondal K, Sharma A. (2013) Reusable electrospun mesoporous ZnO nanofiber mats for photocatalytic degradation of polycyclic aromatic hydrocarbon dyes in wastewater. Journal of Colloid and Interface Science, 394, 208-15.
- 25. Hoffmann MR, Martin ST, Choi W, Bahnemannt DW. (1995) Environmental Applications of Semiconductor Photocatalysis. Chemical Reviews, 95(1), 69-96
- 26. Meng Z, Juan Z. (2008) Wastewater treatment by photocatalytic oxidation of nano-ZnO. Global Environmental Policy in Japan, 12, 1-9.
- 27. Hernandez-Alonso MD, Fresno F, Suareza S, Coronado JM. (2009) Development of alternative photocatalysts to TiO₂: Challenges and opportunities. Energy & Environmental Science, 2(12), 1231-57.
- 28. Vandevivere PC, Bianchi R, Verstraete W. (1998) Treatment and reuse of wastewater from the textile wet-processing industry: Review of emerging technologies. Journal of Chemical Technology and Biotechnology 72(4), 289-302.
- 29. Chan YJ, Chong MF, Law CL, Hassell DG. (2009) A review on anaerobic–aerobic treatment of industrial and municipal wastewater. Chemical Engineering Journal, 155(1-2), 1-18.
- Swaminathan M, Muruganandham M, Sillanpaa M. (2013) Advanced oxidation processes for wastewater treatment. International Journal of Photoenergy, 2013 (683682), 1-3.

- 31. Agustina TE, Ang HM, Vareek VK. (2005) A review of synergistic effect of photocatalysis and ozonation on wastewater treatment. Journal of Photochemistry and Photobiology C: Photochemistry Reviews, 6(4), 264-73.
- Rice RG. (2004) Applications of ozone for industrial wastewater treatment — A review. Ozone-Science & Engineering: The Journal of the International Ozone Association 1997;18(6), 477-515.
- 33. Gogate PR, Pandit AB. (2004) A review of imperative technologies for wastewater treatment II: hybrid methods. Advances in Environmental Research, 8(3-4), 553-97.
- Wojnarovits L, Takacs E. (2008) Irradiation treatment of azo dye containing wastewater: An overview. Radiation Physics and Chemistry 77(3), 225-44
- 35. Azarian G, Mesdaghinia A, Vaezi F, Nabizadeh R, Nematollahi D. (2007) Algae removal by electro-coagulation process, application for treatment of the effluent from an industrial wastewater treatment plant Iranian Journal of Public Health 36(4), 57-64
- Inan H, Dimoglo A, Simsek H, Karpuzcu M. Olive oil mill wastewater treatment by means of electro-coagulation. Separation and Purification Technology 2004;36(1), 23-31
- Arslan I, Balcioglu IA. (2001) Advanced oxidation of raw and biotreated textile industry wastewater with O₃, H₂O₂/UV-C and their sequential application. Journal of Chemical Technology and Biotechnology 76, 53-60.
- Lin SH, Chang CC. (2000) Treatment of landfill leachate by combined electro-Fenton oxidation and sequencing batch reactor method. Water Research 34(17), 4243-9
- Jon P. Scott, Ollis DF. (1995) Integration of chemical and biological oxidation processes for water treatment: Review and recommendations. Environmental Progress 14(2), 88-103.
- 40. Oller I, Malato S, Sánchez-Pérez JA. (2011) Combination of Advanced Oxidation Processes and biological treatments for wastewater decontamination-A review. Science of the Total Environment 409(20), 4141-66.
- 41. Arsene D, Musteret CP, Catrinescu C, Apopei P, Barjoveanu G, Teodosiu C. (2011) Combined oxidation and ultrafiltration processes for the removal of priority organic pollutants from wastewaters. Environmental Engineering and Management Journal 10(12), 1967-76.

- 42. Turchi CS, Ollis DF. (1990) Photocatalytic degradation of organic water contaminants: Mechanisms involving hydroxyl radical attack. Journal of Catalysis 122(1), 178-92
- 43. Veprek-Siska J, Lunak S. (1978) Photocatalytic effects of trace metals evidance against a free radical chain mechanism in sulfite auto oxidation. Reaction Kinetics and Catalysis Letters 8(4), 483-7.
- 44. Ali R, Hassan SH. (2008) Degradation studies on paraquat and malathion using TiO₂/ZnO based photocatalyst. The Malaysian Journal of Analytical Sciences 12(1), 77-87.
- 45. Baruah S, Pal SK, Dutta J. (2012) Nanostructured zinc oxide for water treatment. Nanoscience & Nanotechnology-Asia 2, 90-102.
- 46. Adams LK, Lyon DY, Alvarez PJJ. (2006) Comparative eco-toxicity of nanoscale TiO₂, SiO₂, and ZnO water suspensions. Water Research 40(19), 3527 32.
- 47. Sabin F, Türk T, Vogler A. (1992) Decontamination of industrial waste water by photocatalytic oxidation of organic components: A model study. Zeitschrift Fur Wasser-Und Abwasser-Forschung-Journal for Water and Wastewater Research-Acta Hydrochimica Et Hydrobilogica 25(3), 163-7.
- 48. Mehrvar M, Anderson WA, Moo-Young M. (2001) Photocatalytic degradation of aqueous organic solvents in the presence of hydroxyl radical scavengers. International Journal of Photoenergy 3, 187-91.
- 49. Khezrianjoo S, Revanasiddappa H. (2012) Langmuir-Hinshelwood kinetic expression for the photocatalytic degradation of metanil yellow aqueous solutions by ZnO catalyst. Chemical Sciences Journal 2012(CSJ-85), 1-7.
- 50. Frank SN, Bard AJ. (1977) Heterogeneous photocatalytic oxidation of cyanide ion in aqueous solutions at titanium dioxide powder. Journal of the American Chemical Society 99(1), 303-4.
- 51. Peral J, Domenech X, Ollis2 DF. (1997) Heterogeneous photocatalysis for purification, decontamination and deodorization of air Journal of Chemical Technology and Biotechnology 70(2), 117-40.
- 52. Modestov A, Glezer V, Marjasin I, Lev O. (1997) Photocatalytic degradation of chlorinated phenoxyacetic acids by a new buoyant titania-exfoliated graphite composite photocatalyst. Journal of Physical Chemistry B 101(23), 4623-9.
- 53. Fujishima A, Rao TN, Tryk DA. (2000) Titanium dioxide photocatalysis. Journal of Photochemistry and Photobiology C: Photochemistry Reviews 1, 1–21.

- 54. Sobczynski A, Dobosz A. (2001) Water purification by photocatalysis on semiconductors Polish Journal of Environmental Studies 10(4), 195-205.
- 55. Fan H, Jiang T, Wang L, Wang D, Li H, Wang P, et al. (2012) Effect of BiVO4 crystalline phases on the photoinduced carriers behavior and photocatalytic activity. The Journal of Physical Chemistry 116(3), 2425-30.
- 56. Mills A, Davies RH, Worsley D. (1993) Water purification by semiconductor photocatalysis. Chemical Society Reviews 22(6), 417-25.
- 57. HASHIMOTO K, IRIE H, FUJISHIMA A. (2005) TiO₂ photocatalysis: A historical overview and future prospects Japanese Journal of Applied Physics 44(12), 8269–85.
- Neppolian B, Sakthivel S, Arabindoo B, Palanichamy M, Murugesan V. (1999) Degradation of textile dye by solar light using TiO₂ and ZnO photocatalysts. Journal of Environmental Science and Health, Part A, Toxic/Hazardous Substances and Environmental Engineering 34(9), 1829-38.
- Thiruvenkatachari R, Vigneswaran S, Moon IS. (2008) A review on UV/TiO₂ photocatalytic oxidation process Korean Journal of Chemical Engineering 25(1), 64-72
- 60. Hariharan C. (2006) Photocatalytic degradation of organic contaminants in water by ZnO nanoparticles: Revisited Applied Catalysis A: General 304(1), 55-61.
- 61. Ye M, Zhang Q, Hu Y, Ge J, Lu Z, He L, et al. (2010) Magnetically recoverable core–shell nanocomposites with enhanced photocatalytic activity. Chemistry-A European Journal 16(21), 6243-50.
- 62. Colmenares JC, Luque R, Campelo JM, Colmenares F, Karpinski Z, Romero AA. (2009) Nanostructured photocatalysts and their applications in the photocatalytic transformation of lignocellulosic biomass: An overview. Materials 2(4), 2228-58.
- 63. Anpo M, Shima T, Kodama S, Kubokawa Y. (1987) Photocatalytic hydrogenation of propyne with water on small-particle titania: size quantization effects and reaction intermediates. Journal of Physical Chemistry 91(16), 4305-10.
- 64. Lin H, Huang CP, Li W, Ni C, Shah SI, Tseng Y-H. (2006) Size dependency of nanocrystalline TiO_2 on its optical property and photocatalytic reactivity exemplified by 2-chlorophenol. Applied Catalysis B: Environmental 68(1-2), 1–11.

- Fox MA, Dulay MT. (1993) Heterogeneous photocatalysis. Chemical Reviews 93(1), 341-57.
- 66. Kim DS, Kwak S-Y. (2007) The hydrothermal synthesis of mesoporous TiO₂ with high crystallinity, thermal stability, large surface area, and enhanced photocatalytic activity. Applied Catalysis A: General 323:110-8.
- 67. Wold A. (1993) Photocatalytic properties of titanium dioxide (TiO₂). Chemistry of Materials 5(3), 280-3.
- 68. Nam H-J, Amemiya T, Murabayashi M, Itoh K. (2004) Photocatalytic activity of sol-gel TiO_2 thin films on various kinds of glass substrates: The effects of Na⁺ and primary particle size. Journal of Physical Chemistry B 108(24), 8254-9.
- 69. Wu C-H. (2004) Comparison of azo dye degradation efficiency using UV/single semiconductor and UV/coupled semiconductor systems. Chemosphere 57(7), 601–8.
- Okomoto K, Yamamoto Y, Tanaka H, Itaya A. (1985) Kinetics of heterogeneous photocatalytic decomposition of phenol over anatase TiO₂ powder. Bulletin of the Chemical Society of Japan 58(7), 2023-8.
- Sakthivel S, Neppolian B, Arabindoo B, Palanichamy M, Murugesan V. (2000) TiO₂ catalysed photodegradation of leather dye. Journal of Scientific & Industrial Research 59(7), 556-62.
- Augugliaro V, Palmisano L, Sclafani A, Minero C, Pelizzetti E. (1988) Photocatalytic degradation of phenol in aqueous titanium dioxide dispersions. Toxicology and Environmental Chemistry 16(2), 89-109.
- 73. Jang HD, Kim S-K, Kim S-J. (2001) Effect of particle size and phase composition of titanium dioxide nanoparticles on the photocatalytic properties. Journal of Nanoparticle Research 3(2-3), 141-7.
- Yu JC, Yu J, Ho W, Jiang Z, Zhang L. (2002) Effects of F-doping on the photocatalytic activity and microstructures of nanocrystalline TiO₂ powders. Chemistry of Materials 14(9), 3808-16.
- 75. Testino A, Bellobono IR, Buscaglia V, Canevali C, D'Arienzo M, Polizzi S, et al. (2007) Optimizing the photocatalytic properties of hydrothermal TiO_2 by the control of phase composition and particle morphology. A systematic approach. Journal of the American Chemical Society 129(12), 3564-75.
- 76. Mo S-D, Ching WY. (1995) Electronic and optical properties of three phases of titanium dioxide: Rutile, anatase, and brookite. Physical Review B 51(19), 13023-32.

- 77. Iskandar F, Nandiyanto ABD, Yun KM, Hogan CJ, Jr. KO, Biswas P. (2007) Enhanced photocatalytic performance of brookite TiO₂ macroporous particles prepared by spray drying with colloidal templating. Advanced Materials 19(10), 1408–12.
- Bavykin DV, Friedrich JM, Walsh FC. (2006) Protonated titanates and TiO₂ nanostructured materials: Synthesis, properties, and applications. Advanced Materials 18(21), 2807-24.
- 79. M. Koelsch, S. Cassaignon, J.F. Guillemoles, J.P. Jolivet. (2002) Comparison of optical and electrochemical properties of anatase and brookite TiO_2 synthesized by the sol-gel method Thin Solid Films 403:312–9.
- Morgan BJ, Watson GW. (2010) Intrinsic n-type Defect Formation in TiO₂: A Comparison of Rutile and Anatase from GGA plus U Calculations. Journal of Physical chemistry 114(5), 2321-8.
- 81. Qamar M, Yoon CR, Oh HJ, Lee NH, Park K, Kim DH, et al. (2008) Preparation and photocatalytic activity of nanotubes obtained from titanium dioxide Catalysis Today 131(1-4), 3-14.
- Sclafani A, Herrmann JM. (1996) Comparison of the photoelectronic and photocatalytic activities of various anatase and rutile forms of titania in pure liquid organic phases and in aqueous solutions Journal of Physical Chemistry 100(32), 13655-61.
- 83. Bakardjieva S, Subrt J, Stengl V, Dianez MJ, Sayagues MJ. (2005) Photoactivity of anatase-rutile TiO₂ nanocrystalline mixtures obtained by heat treatment of homogeneously precipitated anatase. Applied catalysis B:Environmental 58(3-4), 193-202.
- 84. Ohno T, Tsubota T, Toyofuku M, Inaba R. (2004) Photocatalytic activity of a TiO₂ photocatalyst doped with C^{4+} and S^{4+} ions having a rutile phase under visible light. Catalysis Letters 98(4), 255-8.
- Colon G, Hidalgo MC, Munuera G, Ferino I, Cutrufello MG, Navio JA. (2006) Structural and surface approach to the enhanced photocatalytic activity of sulfated TiO₂ photocatalyst. Applied Catalysis B: Environmental 63(1-2), 45-59
- Li H, Zhang W, Pan W. (2011) Enhanced photocatalytic activity of electrospun TiO₂ nanofibers with optimal anatase/rutile ratio. Journal of the American Ceramic Society 94(10), 3184-7.
- Ohtani B, Ogawa Y, Nishimoto S-i. (1997) Photocatalytic activity of amorphous-anatase mixture of titanium(IV) oxide particles suspended in aqueous solutions. Journal of Physical Chemistry B 101(19), 3746-52

- Zhou X-T, Ji H-B, Huang X-J. (2012) Photocatalytic degradation of methyl orange over metalloporphyrins supported on TiO₂ degussa P25. Molecules 17(2), 1149-58
- Lydakis-Simantiris N, Riga D, Katsivela E, Mantzavinos D, Xekoukoulotakis NP. (2010) Disinfection of spring water and secondary treated municipal wastewater by TiO₂ photocatalysis. Desalination 250(1), 351-5.
- 90. Giwa A, Nkeonye PO, Bello KA, Kolawole EG. (2012) Solar photocatalytic degradation of reactive yellow 81 and reactive violet 1 in aqueous solution containing semiconductor oxides. International Journal of Applied Science and Technology 2(4), 90-105.
- Chang S-m, Lo P-h, Chang C-t. (2009) Photocatalytic behavior of TOPO-capped TiO₂ nanocrystals for degradation of endocrine disrupting chemicals. Applied Catalysis B: Environmental 91(3-4), 619-27.
- 92. Geng Q, Guo Q, Cao C, Wang L. (2008) Investigation into nanoTiO₂/ACSPCR for decomposition of aqueous hydroquinone. Industrial & Engineering Chemistry Research 47(8), 2561-8.
- Hu X, Li G, Yu JC. (2010) Design, fabrication, and modification of nanostructured semiconductor materials for environmental and energy applications. Langmuir 26(5), 3031-9.
- 94. Reddy BM, Reddy GK, Rao KN, Ganesh I, Ferreira JMF. (2009) Characterization and photocatalytic activity of $TiO_2-M_xO_y$ (M_xO_y 5 SiO_2 , Al_2O_3 , and ZrO_2) mixed oxides synthesized by microwaveinduced solution combustion technique. Journal of Materials Science 44(18), 4874-82
- 95. Zhou W, Liu K, Fu H, Pan K, Zhang L, Wang L, et al. (2008) Multimodal mesoporous TiO₂–ZrO₂ composites with high photocatalytic activity and hydrophilicity. Nanotechnology 19(3), 1-7.
- Tryba B, Morawski AW, Inagaki M. (2003) Application of TiO₂mounted activated carbon to the removal of phenol from water. Applied Catalysis B: Environmental 41(4), 427–33.
- 97. Chen L, Pang X, Yu G, Zhang J. (2010) In-situ coating of MWNTs with solgel TiO₂ nanoparticles. Advanced Materials Letters 1(1), 75-8.
- 98. Zhang P, Shao C, Zhang Z, Zhang M, Mu J, Guo Z, et al. (2011) Core/shell nanofibers of TiO₂@carbon embedded by Ag nanoparticles with enhanced visible photocatalytic activity Journal of Materials Chemistry 21(44), 17746-53.

- 99. Ng YH, Lightcap IV, Goodwin K, Matsumura M, Kamat PV. (2010) To what Extent do graphene scaffolds improve the photovoltaic and photocatalytic response of TiO₂ nanostructured films? Journal of Physical Chemistry Letters 1(15), 2222–7.
- Savoskin MV, Yaroshenko AP, Lazareva NI, Mochalin VN, Mysyk RD. (2006) Using graphite intercalation compounds for producing exfoliated graphite–amorphous carbon–TiO₂ composites. Journal of Physics and Chemistry of Solids 67(5-6), 1205–7.
- Williams G, Seger B, Kamat PV. (2008) TiO₂-graphene nanocomposites. UV-assisted photocatalytic reduction of graphene oxide. ACS Nano 2(7), 1487–91.
- 102. Yin S, Hasegawa H, Maeda D, Ishitsuka M, Sato T. (2004) Synthesis of visible-light-active nanosize rutile titania photocatalyst by low temperature dissolution-reprecipitation process. Journal of Photochemistry and Photobiology A: Chemistry 163(1-2), 1-8.
- 103. Ding X-Z, Liu X-H. (1998) Correlation between anatase-to-rutile transformation and grain growth in nanocrystalline titania powders Journal of Materials Research 13(9), 2556-9
- 104. Bacsa RR, Kiwi J. (1998) Effect of rutile phase on the photocatalytic properties of nanocrystalline titania during the degradation of p-coumaric acid. Applied Catalysis B: Environmental 16(1), 19–29.
- 105. Nakamura R, Nakato Y. (2004) Primary intermediates of oxygen photoevolution reaction on TiO₂ (rutile) particles, revealed by in situ FTIR absorption and photoluminescence measurements. Journal of the American Chemical Society 126(4), 1290–8.
- Becka DD, Siegel RW. (1992) The dissociative adsorption of hydrogen sulfide over nanophase titanium dioxide. Journal of Materials Research 7(10), 2840-5.
- Rio Gd, Rudin A. (1996) Latex particle size and CPVC. Progress in Organic Coatings 28(4), 259–70.
- Zhu K, Neale NR, Miedaner A, Frank AJ. (2007) Enhanced chargecollection efficiencies and light scattering in dye-sensitized solar cells using oriented TiO₂ nanotubes arrays. Nano Letters 7(1), 69-74.
- 109. Miyauchi M, Nakajima A, Fujishima A, Hashimoto K, Watanabe T. (2000) Photoinduced surface reactions on TiO₂ and SrTiO₃ films: Photocatalytic oxidation and photoinduced hydrophilicity. Chemistry of Materials 12(1), 3–5.
- 110. Kumar A, Madaria AR, Zhou C. (2010) Growth of aligned singlecrystalline rutile TiO₂ nanowires on arbitrary substrates and their

application in dye-sensitized solar cells. Journal of Physical Chemistry C 114(17), 7787–92.

- 111. Sabate J, Cervera-March S, Simarro R, Gimenez J. (1990) A comparative study of semiconductor photocatalysts for hydrogen production by visible light using different sacrificial substrates in aqueous media. International Journal of Hydrogen Energy 15(2), 115–24.
- Kansal SK, Singh M, Sud D. (2007) Studies on photodegradation of two commercial dyes in aqueous phase using different photocatalysts. Journal of Hazardous Materials 141(3), 581–90.
- 113. Abe R, Takata T, Sugihara H, Domen K. (2005) Photocatalytic overall water splitting under visible light by TaON and WO₃ with an IO₃ ⁻/I⁻ shuttle redox mediator. Chemical Communications (30), 3829-31.
- 114. Kim J, Lee CW, Choi W. (2010) Platinized WO₃ as an environmental photocatalyst that generates OH radicals under visible light. Environmental Science & Technology 44(17), 6849–54.
- 115. Wang X, Yu JC, Chen Y, Wu L, Fu X. (2006) ZrO₂-modified mesoporous nanocrystalline TiO_{2-x}N_x as efficient visible light photocatalysts. Environmental Science & Technology 40(7), 2369–74.
- 116. Li Y, Xie W, Hu X, Shen G, Zhou X, Xiang Y, et al. (2010) Comparison of dye photodegradation and its coupling with light-toelectricity conversion over TiO_2 and ZnO. Langmuir 26(1), 591–7.
- 117. Sung-Suh HM, Choi JR, Hah HJ, Koo SM, Bae YC. (2004) Comparison of Ag deposition effects on the photocatalytic activity of nanoparticulate TiO_2 under visible and UV light irradiation. Journal of Photochemistry and Photobiology A: Chemistry 163(1-2), 37-44.
- 118. Victor F. Stone J, Davis RJ. (1998) Synthesis, characterization, and photocatalytic activity of titania and niobia mesoporous molecular sieves. Chemistry of Materials 10(5), 1468-74.
- 119. Shankar KS, Raychaudhuri AK. (2005) Fabrication of nanowires of multicomponent oxides:Review of recent advances. Materials Science and Engineering C 25(5-8), 738–51.
- 120. Zhang M, Shao C, Guo Z, Zhang Z, Mu J, Cao T, et al. (2011) Hierarchical nanostructures of copper(II) phthalocyanine on electrospun TiO₂ nanofibers: Controllable solvothermal-fabrication and enhanced visible photocatalytic properties. ACS Applied Materials & Interfaces 3(2), 369–77.

- Jin Y, Wang J, Sun B, Blakesley JC, Greenham NC. (2008) Solutionprocessed ultraviolet photodetectors based on colloidal ZnO nanoparticles. Nano Letters 8(6), 1649-53.
- 122. Gratzel M. (2005) Solar energy conversion by dye-sensitized photovoltaic cells Inorganic Chemistry 44(20), 6841-51.
- 123. Zhang Q, Dandeneau CS, Zhou X, Cao G. (2009) ZnO nanostructures for dye-Sensitized solar cells. Advanced Materials 21(41), 4087-108.
- 124. Ramanathan K, Contreras MA, Perkins CL, Asher S, Hasoon FS, Keane J, et al. (2003) Properties of 19.2% efficiency ZnO/CdS/CuInGaSe₂ thin-film solar cells. Progress in Photovoltaics: Research and Applications 11(4), 225–30.
- 125. Yang JL, An SJ, Park WI, Yi G-C, Choi W. (2004) Photocatalysis using ZnO thin films and nanoneedles grown by metal–organic chemical vapor deposition. Advanced Materials 16(18), 1661–4.
- 126. Devan RS, Patil RA, Lin J-H, Ma Y-R. (2012) One-dimensional metaloxide nanostructures: Recent developments in synthesis, characterization, and applications. Advanced Functional Materials 22(16), 3326–70.
- Fang Lu, Cai W, Zhang Y. (2008) ZnO hierarchical micro/nanoarchitectures: Solvothermal synthesis and structurally enhanced photocatalytic performance. Advanced Functional Materials 18(7), 1047–56.
- Chakrabarti S, Dutta BK. (2004) Photocatalytic degradation of model textile dyes in wastewater using ZnO as semiconductor catalyst. Journal of Hazardous Materials 112(3), 269–78.
- Mathur N, Bhatnagar P, Mohan K, Bakre P, Nagar P, Bijarnia M. (2007) Mutagenicity evaluation of industrial sludge from common effluent treatment plant. Chemosphere 67(6), 1229–35.
- 130. Behnajady MA, Modirshahla N, Hamzavi R. (2006) Kinetic study on photocatalytic degradation of C.I. Acid Yellow 23 by ZnO photocatalyst. Journal of Hazardous Materials 133(1-3), 226–32.
- Daneshvar N, Aber S, Dorraji MSS, Khataee AR, Rasoulifard MH. (2007) Photocatalytic degradation of the insecticide diazinon in the presence of prepared nanocrystalline ZnO powders under irradiation of UV-C light. Separation and Purification Technology 58(1), 91–8.
- Rizzo L. (2011) Bioassays as a tool for evaluating advanced oxidation processes in water and wastewater treatment. Water Research 45(15), 4311–40.

- Gupta VK, Suhas. (2009) Application of low-cost adsorbents for dye removal – A review. Journal of Environmental Management 90(8), 2313–42.
- Hao OJ, Kim H, Chiang P-C. (2000) Decolorization of wastewater. Critical Reviews in Environmental Science and Technology 30(4), 449–505.
- 135. Chen G. (2004) Electrochemical technologies in wastewater treatment. Separation and Purification Technology 38(1), 11–41.
- Li XZ, Liu H. (2003) Photocatalytic oxidation using a new catalystTiO₂ microspherefor water and wastewater treatment. Environmental Science & Technology 37(17), 3989–94.
- 137. R.Gogate P, B.Pandit A. (2004) A review of imperative technologies for wastewater treatment I: oxidation technologies at ambient conditions. Advances in Environmental Research 8(3-4), 501–51.
- 138. Li Q, Mahendra S, Lyon DY, Brunet L, Liga MV, Li D, et al. (2008) Antimicrobial nanomaterials for water disinfection and microbial control: Potential applications and implications. Water Research 42(18), 4591–602.
- 139. Li B, Liu T, Wang Y, Wang Z. (2012) ZnO/graphene-oxide nanocomposite with remarkably enhanced visible-light-driven photocatalytic performance. Journal of Colloid and Interface Science 377(1), 114–21.
- Ueda K, Tabata H, Kawai T. (2001) Magnetic and electric properties of transition-metal-doped ZnO films Applied Physics Letters 79(7), 988-90.
- 141. Garcia SP, Semancik S. (2007) Controlling the morphology of zinc oxide nanorods crystallized from aqueous solutions: The effect of crystal growth modifiers on sspect ratio. Chemistry of Materials 19(16), 4016-22.
- Wang ZL. (2004) Zinc oxide nanostructures: Growth, properties and applications Journal of Physics-Semicondenced Matter 16(25), R829-R58.
- Carraway ER, Hoffman AJ, Hoffmann MR. (1994) Photocatalytic oxidation of organic acids on quantum-sized semiconductor colloids. Environmental Science & Technology 28(5), 786–93.
- 144. Akyol A, Yatmaz HC, Bayramoglu M. (2004) Photocatalytic decolorization of Remazol Red RR in aqueous ZnO suspensions. Applied Catalysis B: Environmental 54(1), 19–24.

- 145. Liu B, Zeng HC. (2003) Hydrothermal synthesis of ZnO nanorods in the diameter regime of 50 nm. Journal of the American Chemical Society 125(15), 4430–1.
- Spanhel L, Anderson MA. (1991) Semiconductor clusters in the sol-gel process: quantized aggregation, gelation, and crystal growth in concentrated zinc oxide colloids. Journal of the American Chemical Society 113(8), 2826–33.
- 147. Liu X, Wu X, Cao H, Chang R. (2004) Growth mechanism and properties of ZnO nanorods synthesized by plasma-enhanced chemical vapor deposition Journal of Applied Physics 95(6), 3141 7.
- 148. Carcia P, McLean R, Reilly M, Jr GN. (2003) Transparent ZnO thinfilm transistor fabricated by rf magnetron sputtering Applied Physics Letters 82(7), 1117 - 9.
- 149. Siddheswaran R, Sankar R, Babu MR, Rathnakumari M, Jayavel R, Murugakoothan P, et al. (2006) Preparation and characterization of ZnO nanofibers by electrospinning. Crystal Research and Technology 41(5), 446–9.
- 150. Fan Z, Lu JG. (2005) Zinc oxide nanostructures: Synthesis and properties. Journal of Nanoscience and Nanotechnology 5(10), 1561-73
- 151. Meulenkamp EA. (1998) Synthesis and growth of ZnO nanoparticles. Journal of Physical Chemistry B 102(29), 5566–72.
- 152. Li Z, Xiong Y, Xie Y. (2003) Selected-control synthesis of ZnO nanowires and nanorods via a PEG-assisted route. Inorganic Chemistry 42(24), 8105–9.
- 153. Meena RC, Pachwarya RB, Meena VK, Arya S. (2009) Degradation of textile dyes Ponceau-S and Sudan IV using recently developed photocatalyst, immobilized resin dowex-11. American Journal of Environmental Sciences 5(3), 444-50.
- 154. Zhou M, Yu J, Cheng B. (2006) Effects of Fe-doping on the photocatalytic activity of mesoporous TiO₂ powders prepared by an ultrasonic method. Journal of Hazardous Materials B 137(3), 1838–47.
- 155. Hagfeldtt A, Gratzel M. (1995) Light-induced redox reactions in nanocrystallinesystems. Chemical Reviews 95(1), 49–68.
- 156. Alhakimi G, Studnicki LH, (2003) Al-Ghazali M. Photocatalytic destruction of potassium hydrogen phthalate using TiO₂ and sunlight: application for the treatment of industrial wastewater. Journal of Photochemistry and Photobiology A: Chemistry 154(2-3), 219–28.
- 157. Choi W, Termin A, Hoffmann MR. (1994) The role of metal ion dopants in quantum-sized TiO₂: Correlation between photoreactivity

and charge carrier recombination dynamics. Journal of Physical Chemistry 98(51), 13669–79.

- 158. Bhatkhande DS, Pangarkar VG, Beenackers AA. (2002) Photocatalytic degradation for environmental applications a review. Journal of Chemical Technology and Biotechnology 77(1), 102–16.
- 159. Bahnemann D. (2004) Photocatalytic water treatment: solar energy applications. Solar Energy 77(5), 445–59.
- 160. Anpo M, Takeuchi M. (2003) The design and development of highly reactive titanium oxide photocatalysts operating under visible light irradiation. Journal of Catalysis 216(1-2), 505–16.
- 161. Han F, Kambala VSR, Srinivasan M, Rajarathnam D, Naidu R. (2009) Tailored titanium dioxide photocatalysts for the degradation of organic dyes in wastewater treatment: A review. Applied Catalysis A: General 359(1-2), 25–40.
- 162. Zang L, Lange C, Abraham I, Storck S, Maier WF, Kisch H. (1998) Amorphous microporous titania modified with platinum(IV) chlorides-A new type of hybrid photocatalyst for visible light detoxification. Journal of Physical Chemistry B 102(52), 10765–71.
- Li D, Haneda H, Hishita S, Ohashi N. (2005) Visible-light-driven N-Fcodoped TiO₂ photocatalysts. 2. Optical characterization, photocatalysis, and potential application to air purification. Chemistry of Materials 17(10), 2596–602.
- 164. Zielińska B, Grzechulska J, Kaleńczuk RJ, Morawski AW. (2003) The pH influence on photocatalytic decomposition of organic dyes over A11 and P25 titanium dioxide. Applied Catalysis B: Environmental 45(4), 293–300.
- Naeem K, Feng O. (2009) Parameters effect on heterogeneous photocatalysed degradation of phenol in aqueous dispersion of TiO₂. Journal of Environmental Sciences 21:527–33.
- Lawless D, Serpone N, Meisel D. (1991) Role of hydroxyl radicals and trapped holes in photocatalysis. A pulse radiolysis study. Journal of Physical Chemistry 95(13), 5166–70.
- Chiron S, Fernandez-Alba A, Rodriguez A, Garcia-Calvo E. (2000) Pesticide chemical oxidation: state-of-the-art. Water Research 34(2), 366–77.
- 168. Zhu X, Castleberry SR, Nanny MA, Butler EC. (2005) Effects of pH and catalyst concentration on photocatalytic oxidation of aqueous ammonia and nitrite in titanium dioxide suspensions Environmental Science & Technology 39(10), 3784–91.

- 169. Kim Tw, Lee M-J. (2010) Effect of pH and temperature for photocatalyticdegradation of organic compound on carbon-coated TiO₂. Journal of Advanced Engineering and Technology 3(2), 193-8.
- Huang CP, Dong C, Tang Z. (1993) Advanced chemical oxidation: Its present role and potential future in hazardous waste treatment. Waste Management 13(5-7), 361–77.
- 171. Konstantinou IK, Albanis TA. (2004) TiO₂-assisted photocatalytic degradation of azo dyes in aqueous solution: kinetic and mechanistic investigations-A review. Applied Catalysis B: Environmental 49(1), 1-14.
- 172. Tang WZ, Zhang Z, An H, Quintana MO, Torres DF. (1997) TiO₂/UV photodegradation of azo dyes in aqueous solutions. Environmental Technology 18(1), 1-12.
- 173. Cheng M, Ma W, Li J, Huang Y, Zhao J. (2004) Visible-light-assisted degradation of dye pollutants over Fe(III)-loaded resin in the presence of H_2O_2 at neutral pH values Environmental Science & Technology 38(5), 1569–75.
- 174. Kanel SR, Al-Abed SR. (2011) Influence of pH on the transport of nanoscale zinc oxidein saturated porous media. Journal of Nanoparticle Research 13, 4035–47.
- 175. Balcioglu IA, Inel Y. (1996) Photocatalytic degradation of organic contaminants in semiconductor suspensions with added H₂O₂. Journal of Environmental Science and Health A 31(1), 123-38.
- 176. Arslan-Alaton I. (2003) A review of the effects of dye-assisting chemicals on advanced oxidation of reactive dyes in wastewater. Coloration Technology 119(6), 345–53.
- 177. Kiriakidou F, Kondarides DI, Verykios XE. (1999) The effect of operational parameters and TiO₂-doping on the photocatalytic degradation of azo-dyes. Catalysis Today 54(1), 119–30.
- 178. Sun J, Qiao L, Sun S, Wang G. (2008) Photocatalytic degradation of Orange G on nitrogen-doped TiO₂ catalysts under visible light and sunlight irradiation. Journal of Hazardous Materials 155(1-2), 312–9.
- 179. Kaise M, Nagai H, Tokuhashi K, Kondo S, Nimura S, Kikuchi O. (1994) Electron spin resonance studies of photocatalytic interface reactions of suspended M/TiO_2 (M = Pt, Pd, Ir, Rh, Os, or Ru) with alcohol and acetic acid in aqueous media. Langmuir 10(5), 1345–7.
- Hussein FH. (2012) Comparison between solar and artificial photocatalytic decolorization of textile industrial wastewater. International Journal of Photoenergy 2012(Article ID 793648), 1-10.

- 181. Takeda N, Torimoto T, Sampath S, Kuwabata S, Yoneyama H. (1995) Effect of inert supports for titanium dioxide loading on enhancement of photodecomposition rate of gaseous propionaldehyde. Journal of Physical Chemistry 99(24), 9986–91.
- 182. Sakthivel S, Neppolian B, Shankar MV, Arabindoo B, Palanichamy M, Murugesan V. (2003) Solar photocatalytic degradation of azo dye: comparison of photocatalytic efficiency of ZnO and TiO₂. Solar Energy Materials and Solar Cells 77(1), 65–82.
- Cassano AE, Alfano OM. (2000) Reaction engineering of suspended solid heterogeneous photocatalytic reactors Catalysis Today 58(2-3), 67–197.
- 184. Lindner M, Theurich J, Bahnemann DW. (1997) Photocatalytic degradation of organic compounds: accelerating the process efficiency. Water Science and Technology 35(4), 79–86.
- 185. Pillai SC, Periyat P, George R, McCormack DE, Seery MK, Hayden H, et al. (2007) Synthesis of high-temperature stable anatase TiO₂ photocatalyst Journal of Physical Chemistry C 111(4), 1605-11
- Paola AD, Bellardita M, Palmisano L. (2013) Brookite, the least known TiO₂ photocatalyst. Catalysts 3, 36-73.
- 187. Lei S, Duan W. (2008) Highly active mixed-phase TiO₂ photocatalysts fabricated at low temperature and the correlation between phase composition and photocatalytic activity. Journal of Environmental Sciences 20, 1263–7.