

REVIEW ARTICLE

Photocatalytic Degradation of Dyes: An Overview

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Abstract: Introduction: The photocatalytic degradation of dyes has been investigated throughout the world irrespective of the level of science in that country. The normal variables considered are the concentration of oxidising species, the concentration of the dye employed, the catalyst used and intensity and source of photons applied for degradation studies. The kinetic data obtained on the decolorization have usually been treated with pseudo first order kinetic expression even there are some exceptions.

Conclusion: This presentation addresses the limitations of the consideration of this topic under these experimental parameters and shows how the study can be directed in future.

Keywords: Dye degradation, Kinetics of degradation, Semi-conductor photo-catalysis, Mechanism of dye degradation, Reactive oxygen species, Pollution abatement

1. INTRODUCTION

Photocatalytic degradation of dyes or other organic pollutants is an intensively pursued research exercise mainly from the last quarter half of the 20th century [1-7]. Nowadays, the synthetic dyes are extensively used in products like clothes, leather accessories, furniture, and plastic products. However, during the dyeing process, nearly 12% of these dyes exclude as waste, and ~ 20% of this wastage enters to the environment [8]. In the dye degradation process, large molecules of dyes get oxidised down into smaller molecules such as water, carbon dioxide, and other mineral byproducts. As stated, the dyeing process does not utilise all the dye molecules, and consequently, a substantial amount of dyes were present in the waste water released from the industry.

Heterogeneous photocatalysis is one of the modern methods widely employed for the degradation or bleaching of the dyes [9]. The process mainly involves the transfer of electrons from the valence band to the conduction band of a semiconductor surface (mostly oxides and sulphides) on illumination with an appropriate wavelength of light. These generated excitons react with oxygen or water produces superoxide anions and hydroxide radicals. These species have the high oxidising power to degrade numerous molecules including industrial dyes. The decontamination processes by these reactive oxygen species and some other species as like various forms of Fenton processes, called in the scientific parlour as Advanced Oxidation Process (AOP). Even though AOP is an important research area in the contemporary literature, we shall restrict the discussion to the semiconductor mediated photodegradation of dyes [10-12].

At this stage, it is necessary to mention the need for a review on this topic. The reasons include the following[13];

- (1) Some research groups are working in this area. Therefore, it is better to assimilate the literature at constant periodicity.
- (2) Photocatalytic degradation of pollutants is one of the methods have some advantages including total deterioration of the pollutant and possibly in the less expensive method.
- (3) The degraded components like water and carbon dioxide are non-toxic.
- (4) The feasibility of deterioration of any pollutant can be *a priori* decided from the numerical values of the oxidation potential of the pollutant and the reagents such as OH[•] radical with standard reduction potential value of around 2 V [14].

Various kinds of dyes are available in the markets as colouring objects. The classification of dye materials is according to the structure of the molecule component, colour and its method of application. The general classification of dyes evolved based on the chromophoric group in the molecular moiety as acridine dyes, azo dyes, anthraquinone dyes, nitro dyes, xanthene dyes and quinine-amine dyes and so on [7]. The studies reported on photocatalytic dye degradation are mainly concerned with the variables like the concentration of the dye, amount of photocatalyst employed, effect of the intensity of the irradiated light, time of irradiation and effect of dissolved oxygen and other species. The kinetics of photocatalytic degradation of dyes is considered to be a pseudo first order reaction with the kinetic data fitted to the equation Heterogeneous photocatalysis is one of the modern methods widely employed for the degradation or bleaching of the dyes [9]. The process mainly involves the

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transfer of electrons from the valence band to the conduction band of a semiconductor surface (mostly oxides and sulphides) on illumination with an appropriate wavelength of light. These generated excitons react with oxygen or water produces superoxide anions and hydroxide radicals. These species have the high oxidising power to degrade numerous molecules including industrial dyes. The decontamination processes by these reactive oxygen species and some other species as like various forms of Fenton processes, called in the scientific parlour as Advanced Oxidation Process (AOP). Even though AOP is an important research area in the contemporary literature, we shall restrict the discussion to the semiconductor mediated photodegradation of dyes [10-12].

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data fit will be considered separately in a subsequent section.

Though extensive studies are reported on the photocatalytic degradation of pollutants in water, certain aspects have not yet received careful attention. The purpose of this presentation is to focus on these issues and to point out what is required in this direction. The literature in this area is increased five times or more during the last 10 years as seen from the data shown in Fig. 1. It is noticed that the number of publications is doubling or more every five-year period. It is therefore natural that people attempt to review the literature at periodic intervals [5-7]. However as said earlier, the research is pursued mostly around oxides (especially TiO₂) and the variables studied are mainly the same, whether it is required or not.

Before we embark on the limitations of the studies so far reported, it is necessary to review the available literature though not comprehensively but representatively briefly. A few publications from the literature are summarised in Table 1. The majority of the studies reported in literature deal with the effect on degradation activity on variables like the amount of the catalyst, the concentration of the dye employed, pH, effect of the radiation source and time of irradiation and also the effect of dissolved oxygen and others. The kinetics of degradation of dyes on most of the catalyst systems studied follows first order [15].

Conventional chemical, physical and biological processes have been extensively employed for treating waste water containing dye molecules. These methods have the following disadvantages like high cost, the requirement of high energy, generation of secondary pollutants in the treatment process. The Advanced Oxidation Process (AOP) has received considerable attention in recent times for the decomposition of organic dyes [16].

2. THE LITERATURE SO FAR

This is an area of research which is carried out throughout the world unlike other areas of science. Research in particular areas of science is confined to ascertain regions of the world, but the degradation of dyes has been studied in almost all the countries and regions including almost all the developing countries around the world. This is reflected in the data assembled in Table 1. Scientifically the process involved in the degradation of dyes can be pictorially represented as shown in Fig. 2.

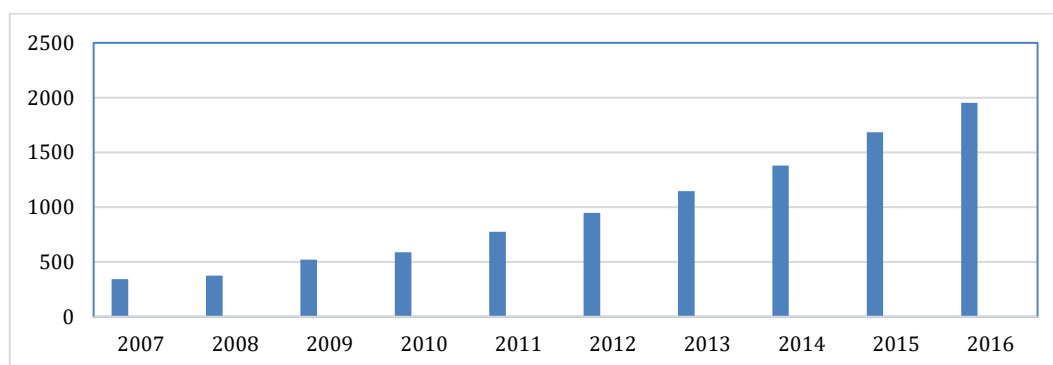


Fig. (1). Number of publications falling under the category of Photo-catalytic degradation of dyes (Source: Web of Science).

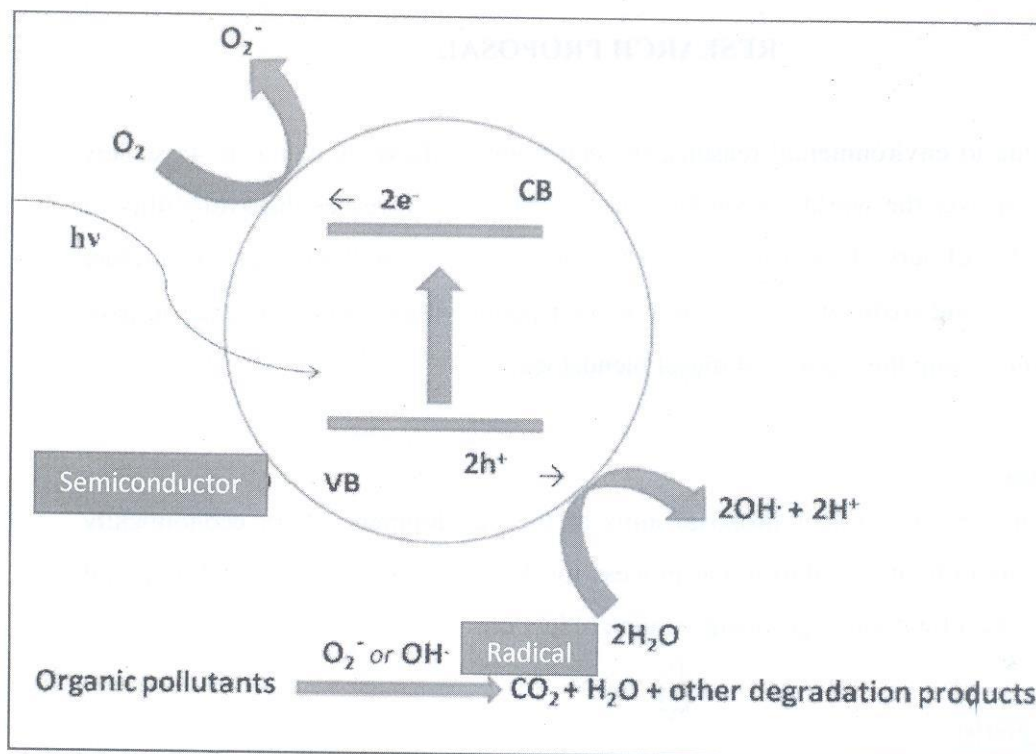


Fig. (2). Pictorial representation of the process taking place in the photocatalytic degradation of dyes on semiconductor surfaces.

Most of the photocatalytic dye degradation studies reported have been with Titanium dioxide as a photocatalyst. However, the major disadvantage of TiO₂ that it absorbs only in the UV region since it has a band gap of around 3.2 eV. Among the different phases of TiO₂, anatase form of TiO₂ is mostly employed due to its higher photon absorption characteristics. It is clear that the phase composition of TiO₂ has a role to play in degradation of dyes. Among the most prominent phases of TiO₂ namely Anatase, rutile and Brookite, the first two phases are most studied systems as seen from the data given in Table.1. The position of oxygen ions on the anatase surface is in a triangular arrangement which allows significant absorption of organic molecules, whereas, the orientation of titanium ions in the anatase phase creates a favourable reaction condition with the absorbed organic pollutants [17-24]. Interestingly, these favourable structural arrangements of oxygen and titanium ions are not present in the rutile phase. It is also believed that pure anatase with a small proportion of rutile phase is conducive for mesoporosity and thus favourable for dye adsorption [25-36].

Mechanistically, the photon excites an electron from the valence band to the conduction band, and the excitons (free electron in the conduction band and hole in the valence band) generate radical species which is responsible for the degradation of organic dyes to carbon dioxide and water and other degradation species.

Even though the large surface area is recommended for the effective degradation of the dye, the adsorption may precede the degradation, and this can affect the interpretation of the kinetics of degradation of dye. This aspect will be taken up subsequently. In the case of Degussa P-25 TiO₂, it is a mixture of 80% anatase and 20% rutile phase, and this combination alone makes this system active and in most

cases used as a standard for comparison, and it is believed to be core shell model system.

Among the various waste water treatment procedures, dye removal has occupied a prominent place. Because of aesthetic and environmental concerns, the degradation of dyes in the effluent water of textile dyeing and finishing industry has been most significant [37]. The semiconductors especially TiO₂ and ZnO are employed as nanorods, nanospheres, thin porous films, nanofibers and nanowires or supported on polymeric films [40]. These systems exhibit high activity, low cost and environmentally acceptable [41-43].

Apart from TiO₂ and ZnO, various other semiconducting systems like CdS, ZrO₂ and WO₃ have been employed in the photocatalytic degradation of dyes. These studies and other reports on ternary oxides are included in the listing in Table.1. The drawback of most of these systems like TiO₂ is the high value of band gap, and they require UV photon sources to be able to decolourize waste-water.

A typical pictorial representation of photodegradation of dyes on a complex photo-catalyst K₆Ta_{10.8}O₃₀ is shown in Fig. 3. In general, on other layered systems and composite catalysts also the photodegradation can take place, and the pictorial representation of this process is shown in Fig. 4(a) and (b).

Brookite is another phase of titanium dioxide which has been used as a photocatalyst for the degradation of dyes [54].

Most of the studies on the photocatalytic degradation of dyes are monitored by decolourization or Chemical Oxygen Demand (COD) measurements though it is desirable to follow the concentration of products formed and elucidate the

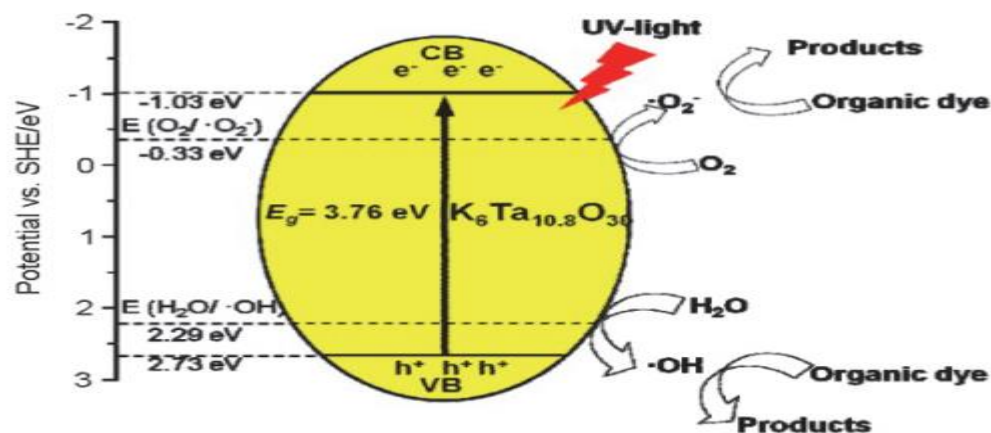


Fig. (3). Representation of energy levels and species responsible for the photodegradation of dyes on typical complex oxide potassium tantalate [Reproduced from Ref.146].

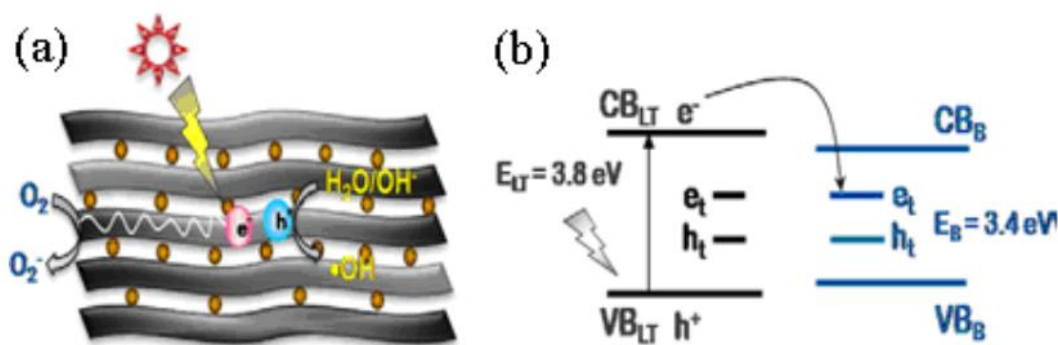


Fig. (4). Diffusion behaviour of the charge carrier in layered nanosheets and (b) transport pathway of the excited electron in the mixed photocatalysts [55].

various pathways in which these products are formed. These type of studies is limited in the literature [56]. It is essential to make sure that the degradation products are not toxic either to the material to be dyed or to living beings. The influence of parameters employed for dyeing process is not established as purely safe. The optimisation of parameters and identifying suitable photocatalyst may be necessary for successful implementation of this technique for purifying waste water in the dyeing technology

The percentage degradation of dyes in waste water improved with increasing intensity of exposed light. With high-intensity irradiation, the recombination may not be significant, but when the intensity is low, the recombination of the electron hole formed predominates. The photocatalytic activity depends on the thermal history of the semiconductor and the chemical nature of the semiconductor. The choice of the semiconducting systems is based on parameters like the physical form of the semiconductor and their stability under the reaction conditions. Environmentally acceptable, cost effectiveness, less toxicity and in all these counts titanium dioxides appears to be the common (best) choice. The following order has been proposed in most of the published literature Degussa P-25 > TiO₂ (Anatase) > TiO₂ (Rutile) for the comparison of various photocatalytic systems towards actual wastewater treatment. However, the amount of catalyst employed depends on the chemical nature of the semiconductor.

The photocatalytic activity can be altered with modification of the semiconductor. The modification can be with various aims like shifting the irradiation wavelength to the visible region and also coupling semiconductors for efficient use of the excited electron-hole pair. Recently g-carbon nitride (g-C₃N₄) has been modified with calcium chloride, and the mechanism of degradation of Rhodamine B dye itself is modified. The proposed schematic diagram is shown in Fig. 5.

The valence band level in the modified system is shifted to more positive value and thus enhances the oxidation ability. Simultaneously the dye is also photoexcited and transfers the electrons to the conduction band of the modified g-C₃N₄. This route predominates when visible light is employed.

Apart from these inherent modifications to the semiconductors, (so called doping), the coupling of semiconductors have also been tried for shifting the wavelength to the visible region, and this is known as a Z-scheme process.

2.1. Mechanism of Photo-catalytic Degradation of Dyes

It has been stated that radical species generated during photoexcitation of the semiconductor is responsible for the degradation of dyes. The essential steps involved can be visualised (in a general sense) in the following steps [42-44].

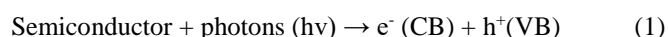


Table 1. Representative literature data on the photocatalytic degradation of dyes

Catalyst Systems Studied	Dyes Employed	Comments	Refs.
Graphene – gold Nano composite (GOR/Au)	Rhodamine B Methylene blue Orange H	The rate of degradation of methylene blue is greater than Rhodamine B in the presence of visible light even though the redox potential is highest among these three dyes. Adsorption of the dye on to the surface identified as the reason for the photoactivity.	[25]
Nanocrystalline anatase and rutile TiO ₂	Acetophenone Nitrobenzene Methylene blue Malachite green	The activity of Anatase is higher than that observed with Rutile. The reason for this difference is not indicated in this communication	[26]
TiO ₂ , ZnO, SnO ₂	Crystal Violet Methyl Red	ZnO shows better photoactivity than Degussa P-25 and silver deposited ZnO increases the photocatalytic activity by 20%	[27]
Mg ²⁺ -TiO ₂	Methyl Orange	The catalyst has better activity than the un-doped TiO ₂ - Dye sensitization and injection of the excited electron is considered as the cause	[28]
TiO ₂ Impregnated ZSM-5 (TiO ₂ -ZSM = 0.15:1)	Reactive Black-5	The system shows high adsorption capacity and degradation activity.	[29]
ZnO-nanoflowers	Methyl Orange Congo Red Eosin B Chicago Sky Blue	The catalyst prepared from asymmetric Zn(ii)dimeric complex showed good photocatalytic activity towards methyl orange compared to other dyes	[30]
ZnO Nano powder	Rhodamine B	95% degradation of the dye was observed under solar light irradiation	[31]
TiO ₂	Methyl Orange Methylene Blue	The photocatalytic activity is found to be greater in the presence of solar light than in UV.	[32]
Nano-sized GdCoO ₄	Rhodamine B Rhodamine Blue(RBL) Orange G(OG) Remazol Brilliant Blue (RBBR)	The catalyst (3nm) is more efficient than P-25. Size dependence is shown. The intermediates in both GdCoO ₄ and P-25 are the same	[33]
TiO ₂	Methylene Blue Methyl Orange Congo Red	The size and Phase (Anatase) are important. Adsorption of the dye on the catalyst surface is also important (Freundlich isotherm)	[34]
TiO ₂	Indigo Indigo Carmine	Complete mineralisation of the dyes Irradiation with visible light only produced colour removal	[36]
TiO ₂ immobilised on polyvinyl alcohol (PVA) or polyacrylamide (PA)	Methylene Blue, Anthraquinone, Remazol Brilliant Blue R (RBBR), Reactive Orange (RO16).	TiO ₂ loaded on PVA appears to be better than that loaded on PA	[35]
Nanostructured TiO ₂	Mono, di and tri azo class of dyes. Categories of indigoid, anthraquinone triaryl methane and xanthen dyes	Degradation depends on the chemical structure of the dye, the nature of functional groups. Mono-azo dyes degrade faster than anthraquinone dyes. The presence of nitrite group promotes the degradation activity.	[4]
High surface area TiO ₂	Methylene Blue Congo Red	Sol-gel method preparation of TiO ₂ is suitable for degradation of Dyes. Freundlich Isotherm is employed.	[2]
N-doped TiO ₂	Methylene Blue Methyl Orange	Visible light source was employed and depends on nitrogen content of the catalyst	[1]

Nanometer sized TiO ₂	Acid Orange 10(AO10) Acid Red 14 14ARI14)	The azo and Sulphonate groups determining factor for degradation	[37]
SiO ₂ nanoparticle doped with Ag and Au	Methyl Red	OH radical produced initiates and also sustains the degradation of the dye	[38]
Titanium dioxide	Emerald Green	Degradation rate constant depends on pH	[39]
ZnO and TiO ₂	Rhodamine B Methylene Blue Acridine Orange	ZnO dissolves as Zn(OH) ₂ and hence shows lower activity as compared to TiO ₂	[40]
TiO ₂ (UV/Solar/pH)	Procion Yellow	TiO ₂ in the presence of solar irradiation gives better degradation.	[41]
TiO ₂	Reactive Red 2	The degradation in the presence of H ₂ O ₂ and persulphate ion were studied.	[42]
Thermally activated ZnO	Congo Red	Pseudo second order Kinetics was observed	[43]
Sol-gel TiO ₂ films	Lissamine Green B	The TiO ₂ film prepared in Polyethylene glycol shows better photoactivity than the clean TiO ₂ .	[44]
ZnO	Methylene Blue	Actual industrial waste water is used for the experiments.	[45]
Ag-TiO ₂ core shell particle	Reactive Blue 220	The core shell system shows better photoactivity under solar light	[46]
Anatase Nano-TiO ₂	Reactive Blue 4 (anthraquinone dye)	In the presence of externally added H ₂ O ₂ , the dye degradation gets increased.	[47]
TiO ₂ /ZnO Photo catalyst	Methylene Blue	ZnO appeared to be better than Pure TiO ₂	[48]
P160 TiO ₂	C! Basic Yellow - 28	Better degradation in weak acidic conditions, the addition of carbonate ion increased the degradation.	[49]
Ferrihydrite modified Diatomite with TiO ₂ /UV	Vat Green 03	A composite catalyst with P-25 with co-adsorbent removed colour over 98%	[50]
Orthorhombic WO ₃	AO7 dye	Phenol, humic acid and EDTA inhibited but oxalic acid increased the decolourisation of the dye.	[51]
Fe ³⁺ /C/S/-TiO ₂	Mono and Di-azo dyes	Mono azo dye degrades faster than diazo dyes under visible light.	[52]
Ni doped TiO ₂	Malachite Green	Hydroxyl ion as the oxidising species	[53]
TiO ₂	Solo phenyl Red 3BL	The concentration of OH [•] and O [•] radical determines the degradation rate.	[54]
TiO ₂	Mono Azo Orange 7 (AO&) Reactive Green 19 (RG19)	Mono azo dye (AO7) degrades faster than binary azo dye (RG19) under solar light.	[55]
TiO ₂	Azo dye and disperse dye	A modelling exercise on governing parameters.	[56]
TiO ₂	Methyl Orange Methylene Blue	Degradation under UV irradiation	[57]
TiO ₂ Photo-catalyst	Indigo Carmine dye	UV irradiation optimum conditions pH =4 and dye concentration 25 ppm 98% colour removal	[58]
ZnO photocatalyst	Methylene Blue	The basic solution is better for the degradation reaction.	[59]
TiO ₂ Photo-catalyst	Methylene Blue	The basic medium is better for the degradation	[60]
Carbon doped TiO ₂	Amido Black-10B	Active oxygenated species is responsible for decolourization.	[61]
ZnO photocatalyst	Direct Red-31 (DR-31) dye	Effect of annealing temperature (500-800C)- UV irradiation	[62]
Sol-gel TiO ₂ films	Methyl orange, Congo Red	TiO ₂ films with dip coating with Polyethylene glycol shows better activity in 254nm than 365 nm.	[63]
Undoped and Fe doped CeO ₂	Methyl Orange	1.5 % doping of Fe ³⁺ was optimal	[64]

Immobilized TiO ₂	Methylene Blue	Deposition of Photosensitive hydroxides decreased the activity	[65]
Ni/MgFe ₂ O ₄	Malachite Green	Visible light active	[66]
TiO ₂	Methyl Orange	Superoxide anion radical Polytetrafluoroethylene-A1 based triboelectric nanogenerator (TENG) assisted the process	[67]
Crosslinked Chitosan/nano CdS	Congo Red	The degradation is better in acidic medium. The presence of NO ₃ ⁻ accelerated, Br ⁻ , Cl ⁻ and SO ₄ ²⁻ inhibited decolorization of the dye.	[68]
TiO ₂ /UV	Methylene Blue	Mineralization of carbon, nitrogen and sulphur into CO ₂ , NH ₄ ⁺ , SO ₃ ²⁻	[69]
Cu impregnated P-25	Azo dye Orange II	Cu Impregnated TiO ₂ is better than H ₂ O ₂ /UV homogeneous reaction.	[70]
Ag-Ni/TiO ₂ synthesised by gamma irradiation	Methyl Red	Bimetallic co-doped is better than bare TiO ₂	[71]
Cr doped TiO ₂	Methylene Blue Congo Red	Cr doped promoted Anatase to Rutile phase transition	[72]
ZnS Quantum dots doped with Au and Ag	Methylene Blue	Metal loading favours degradation; accounted in terms increased the life time of charge carriers. Up to electronic characteristics and isoelectric point need to be considered in proposing photocatalyst.	[73]
Mesoporous CeO ₂	Rhodamine B	Hydroxyl radicals are the active species	[74]
ZnS	Rose Bengal	Hydroxyl radicals are shown as the active species	[69]
C-TiO ₂ films	Azorubine	Photo-degradation and adsorption dual effect is the reason for better decolorization.	[75]
La-Y/TiO ₂	Methylene Blue	Optimum catalyst dose 4 g/L	[76]
Ag-TiO ₂	Direct Red 23	Optimum catalyst dose 3 g/L	[77]
ZnO	Remazol Brilliant Blue dye (RBB)	The degradation follows first order kinetics	[78]
TiO ₂ Degussa P-25	2,4-dimethylphenol, 2,4-dichlorophenol, 2-chlorophenol and phenol	pH 5 was found to be suitable for the degradation reaction.	[79]
ZnO	Crystal Violet	high specific surface area (56.8 m ² /g), high crystallinity and better optical property are responsible for the better activity of ZnO nanonails.	[80]
In/ZnO nano particles	Methylene Blue	Indium is well dispersed on ZnO	[81]
TiO ₂ Degussa P25 and ZnO	Methylene Blue	ZnO shows better activity in visible light than TiO ₂	[82]
TiO ₂ nano particles	Methylene Blue	Basic medium is better for the degradation	[83]
ZnO	Reactive Blue	Reactor design and optimum time	[84]
Magnetite+H ₂ O ₂ +UV	Methylene Blue	Process parameter optimization	[85]
Bi ₂₄ O ₃₁ Cl ₁₀	Rhodamine B	compatible energy levels and high electronic mobility	[86]
BiOI	Rhodamine B anionic reactive blue KN-R	h ⁺ is the dominant species for the degradation of dyes.	[87]
TiO ₂	Alizarin yellow	The presence of Cl ⁻ , SO ₄ ²⁻ inhibits dye removal and it also depends on the TiO ₂ source.	[88]
TiO ₂ , ZnO	Polycyclic aromatic hydrocarbons (AH)	Surface to volume ratio appears to be relevant	[89]
ZnS doped with Mn	Malachite green	UV/ZnS, UV/ZnS/H ₂ O ₂ , UV/doped ZnS systems studied	[90]
TiO ₂ and Cu-doped TiO ₂	reactive blue 4, reactive orange 30, reactive red 120 and reactive black 5	Cu-doped TiO ₂ nanoparticles are very effective in degrading the dye pollutants	[91]

Mn ₃ O ₄ nano particles	amido black 10B	peroxomonosulfate (PMS), peroxydisulfate (PDS) and hydrogen peroxide (HP) enhanced degradation	[92]
Photo-Fenton system	Reactive orange M2R dye	Acidic pH favours, a mechanism is proposed	[93]
TiO ₂ catalyst with a shallow level of Pt	Phenol	Eosin Y sensitised TiO ₂	[94]
TiO ₂	Methylene Blue	p-n junction heterostructure CuO-TiO ₂ enhance photoactivity	[95]
TiO ₂ coated Cotton fabric	amaranth dye	prepared fabric showed enhanced dye degradation capabilities	[96]
titanium dioxide TiO ₂ and zinc phthalocyanine (ZnPc)	4-Nitrophenol	Efficiently degrade nitrophenol	[97]
Silver phosphate	Methylene Blue	Visible-light-driven photodegradation of dye pollutants	[98]
CeCrO ₃	Fast Green dye	First order kinetics,	[99]
ZnO	Acid Green 25	Both acidic and basic medium	[100]
Anatase TiO ₂	Methylene Blue Phenol	pH = 6.4 is optimum	[101]
CeO ₂ -ZnO	Methylene Blue 4'-(1-methyl-benzimidazolyl-2)-phenylazo-2''-(8''-amino-1''-hydroxy-3''',6''-disulphonic)-naphthalene acid	50-80 nm with large defects	[102]
Al ₂ O ₃ -TiO ₂ and ZrO ₂ -TiO ₂ Nanocomposites	Methylene Blue Rhodamine B Methyl Orange	both the composites degrade methylene blue and rhodamine B effectively under UV-A light, and the photodegradation of methyl orange is found to be slow	[103]
MgO	Methylene Blue	Over 90% degradation	[104]
TiO ₂	Acid Orange 67	UV light source is better in comparison to Visible light.	[105]
TiO ₂ on Polyethylene film	Crystal Violet Methylene Blue Basic Fuchsin	Sun light degradation. Undergraduate experiment	[106]
Mo doped TiO ₂	Toluidine blue-o	Degradation of the dye follows pseudo-first order kinetics	[107]
Copper Ferrite	Methylene blue Glycerol	Degradation of glycerol is not efficient and increased in the presence of added H ₂ O ₂ .	[108]
TiO ₂ as photo-catalyst	Tartrazine (azo dye)	Influence of addition of other salts studied	[109]
Ni _{0.6} Co _{0.4} Fe ₂ O ₄	Congo Red	Photo-catalytic degradation maximum at pH 3	[110]
Zn-TiO ₂	Direct Blue 71 dye	Zn Doped system is better than bare TiO ₂	[111]
Ag modified ZnO	Reactive Orange 16	Ag modified system was better than pure ZnO	[112]
TiO ₂	Reactive Orange 16 Dye (RO16)	Effect of the amount of TiO ₂ studied	[113]
ZnO-CuO	Reactive black5 (RB5)	This system is a suitable technique for degradation of dyes and environmental pollution from effluents.	[114]
TiO ₂ on polyethylene glycol	Methyl Orange Congo Red	Under UV irradiation higher efficiency observed	[115]
g-C ₃ N ₄ thermally Modified with Calcium Chloride	Rhodamine B	The photo-generated hole and the superoxide radical is the main active species in the degradation process. 50 times more active than unmodified system	[116]

CdO/TiO ₂ coupled semiconductor	Reactive Orange 4 (RO 4)	best photocatalytic activity in the degradation of RO 4 compared with bare TiO ₂	[117]
ZnO	Remazol Brilliant Blue R, Remazol Black B, Reactive Blue 221 and Reactive Blue 222	A synergistic effect in the coupled TiO ₂ -ZnO system was not observed	[118]
CdS/SL g-C ₃ N ₄) SL= Single Layer)	Rhodamine B	visible-light-responsive and environmentally friendly photocatalyst for the degradation of dye	[119]
BiOCl	Rhodamine B and other dyes	Visible light degradation may be complicated. The use of multitude of dyes is necessary to assess the degradation activity	[120]
Cr doped ZnS	Methyl Orange	Visible light is better than UV	[121]
Nano TiO ₂ (C-Fe doped)	C.I. Basic blue 9 C.I. Acid orange 52	Real waste water treatment	[122]
CeO ₂ -SnO ₂	Direct Black 38	Activity is comparable with TiO ₂ -P25	[123]
Z-scheme SnO _{2-x} g-C ₃ N ₄ composite	Rhodamine B	Z-scheme mechanism to enhance photo-degradation activity	[124]
BiOCl-Au-CdS	Methyl Red Rhodamine B	Z-scheme BiOCl-Au-CdS exhibited excellent sunlight-driven photocatalytic activity toward the degradations of organic dyes and antibiotics	[125]
TiO ₂ -ZnO	RB 21 dye	UV photoreactor and TiO ₂ is the best	[126]
CaO	indigo carmine dye	pH 9 was suitable	[127]
g-C ₃ N ₄ /oxygen vacancy-rich zinc oxide	Methyl Orange	deactivated after five cycles of methyl orange degradation	[128]
CoFe ₂ O ₄ /C ₃ N ₄ hybrid	Rhodamine B	Typical Z-scheme system in environmental remediation	[129]
α -Bi ₄ V ₂ O ₁₁ ; γ -Bi ₄ V ₂ O ₁₁	Rhodamine B Methylene Blue	Surface to Volume ratio is responsible	[130]
BiVO ₄ -rGO	Rhodamine B	Better than pure BiVO ₄ and P-25	[131]
Flower like N-doped MoS ₂	Rhodamine B	27 times better than bare MoS ₂ and 7 times better than P-25	[132]
H ₃ PW ₁₂ O ₄₀ /SiO ₂	Rhodamine B	under simulated natural light irradiation	[133]
SrTiO ₃	Methylene Blue Rhodamine Methyl Orange	Non-selective process	[134]
CuO/Ag ₃ AsO ₄ /GO	Phenol	Photo-stability and reusability	[135]
TiO ₂ /diatomite	Rhodamine B, Methyl orange, Methylene blue	wastewater treatment -good photocatalytic property and reusability.	[136]
Cr(VI) using Ag/TiO ₂	4-chlorophenol	stability and reusability of catalysts	[137]
PbCrO ₄ /TiO ₂	Rhodamine B	good visible light-sensitive photo-catalyst for removing Rh B	[138]
WO ₃ /SnNb ₂ O ₆	Rhodamine B	Z-scheme charge transfer mechanism was proposed for the elimination of organic contaminants under irradiation of visible light.	[139]
ZnO	Acid Red 27	H ₂ O ₂ , K ₂ S ₂ O ₈ , KBrO ₃ due to concentration increases the rate	[140]
CuS	methylene blue, rhodamine B, eosin Y and Congo red	photodegradation rates of dyes usually follow pseudo-first-order kinetics for degradation	[141]
Cobalt Hexacyanoferrate(II)	Neutral Red dye	Degradation under UV light and photo-catalyst	[142]
N-doped ZnO	Azure A	N-doped zinc oxide has been used as an effective catalyst for carrying	[143]

		out number of chemical reactions	
Al ₂ O ₃ -TiO ₂ , ZrO ₂ -TiO ₂	methylene Blue Rhodamine B	Methylene blue degradation is slow Visible light degradation of rhodamine B	[144]
titanium dioxide (TiO ₂), zinc oxide (ZnO), stannic oxide (SnO ₂), zinc sulphide (ZnS), cadmium sulphide (CdS)	Methyl Orange (MO), Rhodamine 6G (R6G)	ZnO/solar light was observed to be better than ZnO/UV system	[145]
K ₆ Ta _{10.8} O ₃₀	ARG dye	the high photocatalytic activity of the degradation of ARG dye under UV	[146]
TiO ₂ (Brookite)	Rhodamine	Less active than P-25	[147]
TiO ₂ (Brookite +rutile mixture)	Orange dye	Samples annealed at different temperatures	[148]
Brookite nanoflowers	Methyl Orange	More active than Anatase	[149, 150]
Anatase and rich brookite rich films	Acid orange 7 4-chlorophenol	Both exhibited same activity	[151]
TiO ₂	Methylene blue Chromium (VI)	TiO ₂ /UV, system is very efficient compared with different natural and artificial adsorbent	[152]

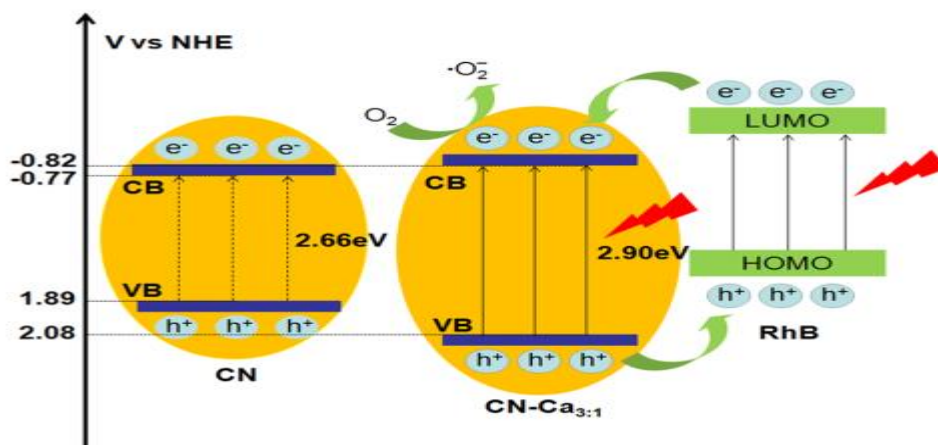
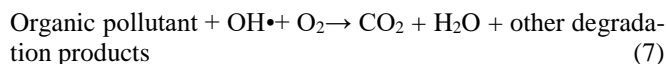


Fig. (5). Energy level diagram of CN and CN modified with CaCl₂ and how the degradation activity of Rhodamine B is enhanced with modification of CN [reproduced from ref. 153].



The mechanism given is a general one and depending on the experimental conditions, additional steps can be included.

The pictorial representation of this process is shown in Fig. 2. The excited electron and hole in the semiconductor are responsible for the degradation of the dye. A variety of semiconductors have been employed, and most of them are used in the nano-state due to increased surface area and also due to favourable quantum size effect [45-48].

TiO₂ in various forms with metal and non-metal doping have been employed for the degradation of a variety of dyes owing to its stability, degradation capability, and also non-toxic nature [49,50]. However, the possible experimental variables including the wavelength of the light to be used and separation technology of the solid in treatment process restrict the employment of TiO₂ for commercial dye degradation process. More advanced level research is at present required to find a suitable alternative to TiO₂ for this application. Other than TiO₂ the other system that is mostly em-

ployed is ZnO and other semiconducting oxides as stated above [51]. However, it is necessary to state the reasons for the preponderance of studies employing TiO₂. Though this is an accepted observation, the reasons for this choice are not explicitly clear in the literature. It is believed that the absorption coefficient (ϵ) of TiO₂ is high. The scattering capacity of the semiconductor has not yet been fully taken into account though it is known that most of the surfaces depending on the value of roughness factor, the scattering intensity can vary. This may also apply to TiO₂. The value of the absorption coefficient depends on various physical properties like the surface roughness, a particle size in addition to the inherent occupancy of the valence electrons and the wave functions of the occupied states.

2.2. Experimental Variables Studied

In addition to the chemical nature of the semiconductor employed, the wave length of irradiation employed based on the band gap of the semiconductor, the effect on the degradation of dyes on a number of other experimental variables has been studied. Typical semiconductors studied and the band gap values of each of them are assembled in Table.2. The photon source employed and intensity of the radiation used is not always reported or monitored. Energy efficient light emitting diodes have also been used, and these studies are reviewed in ref.55. The reaction temperature, the intensity of light and the source of the irradiation, the particle size of the catalyst used, BET-surface area of the solid catalyst used and different mineral forms of the semiconductor influence the degradation rate [56].

2.2.1. Effect of pH on the Photocatalytic Degradation of Dyes

As seen from Table 1, each of the degradation studies is efficient at a particular pH. The reason for this observation is the change in the value of the oxidation potential (approximately 59 mV per pH) of the species involved in the experimental system studied. Since the oxidation potential of the hole and reduction power of the electron generated due to irradiation are dependent on the positions of the top of the valence band and bottom of the conduction band and these are critical for the degradation of dyes on semiconducting systems employed.

The effect of pH on the efficiency of dye photodegradation process has to be associated with multiple roles [57]. The adsorption characteristics of the dye also change pH and the surface characteristics of the semiconductor [58]. This aspect has been already considered in detail in another publication [59]. Hydroxyl radicals can be formed by the reaction of hydroxide ions with positive holes. The positive holes are considered as the major oxidation species at low pH, whereas hydroxyl radicals are considered as the predominant species at neutral or high pH levels [60]. In alkaline solution, •OH are easier to be generated by oxidising hydroxide ions available on the semiconductor surface, thus account for the efficiency of the process is logically enhanced [60]. Similar results have also been reported in the photo-degradation of other dyes [61-63]. It has been postulated that the dyes will be charged as a function of pH and bromo-cresol purple dye degradation was better in the acidic medium than in alkaline medium. There

Table 2. Typical Semiconductors [Refer to Table 1] used for Photo-catalytic Degradation of Dyes and the Band gap (eV) Values of these Materials.

Semiconductors Studied for Photodegradation of Dyes	Band Gap Values (eV) (Wave-length [nm] of Irradiation)
TiO ₂ (Anatase form)	3.2(387)
TiO ₂ (Rutile form)	3.0 (415)
TiO ₂ (Brookite form)	3.14(395)
ZnO	3.36(370)
WO ₃	2.76(450)
CdS	2.42(515)
CuO	1.2 (1035)
Cu ₂ O	2.2 (565)
MgO	5.90
Mn ₃ O ₄	3.28(380)
ZnS	3.6(345)
CeO ₂	3.19(390)
Fe ₂ O ₃	2.3(540)
Fe ₃ O ₄	2.25(550)
ZrO ₂	3.87(320)
g-C ₃ N ₄	2.66(465)
Ag ₂ O	1.4(885)
SrTiO ₃	3.25(380)
Bi ₂ WO ₆	3.13(395)
BaTiO ₃	3.30(375)
Bi ₂ O ₃	2.80(440)
CdO	2.20(560)
CoO	2.01(620)
Cr ₂ O ₃	3.50(355)
HgO	1.90(650)
In ₂ O ₃	2.80(440)
MnO	3.60(345)
Nb ₂ O ₅	3.40(365)
NiO	3.50(355)
PbO	2.80(440)
PdO	1.00(1240)
Sb ₂ O ₃	3.00(415)
SnO	4.20(295)
SnO ₂	3.50(355)
V ₂ O ₅	2.80(440)
K ₆ Ta _{10.8} O ₃₀	3.76(330)

are also views that the charged state cannot be due to change in pH [64]. The route by which photo-degradation takes place depends on the products formed as these product mol-

ecules will be adsorbed on the surface of the semiconductor surface and thereby can alter its electronic and active site configuration. It was reported that in photo-catalytic degradation the extent of adsorption on unmodified TiO₂ is greater for dyes with a positive charge (cationic) than for those with a negative charge (anionic). [66]. Therefore, it is clear that the nature of the particular dye and pH have a profound effect on photocatalytic activity. [67-72]. Azo dyes are positively charged at low pH < 6.8 and higher pH, the dye is negatively charged, and so the adsorption of the dye on the semiconductor surface is affected. It is clear that the waste water treatment must take into account two factors

namely the pH of the effluent is not neutral, and surface properties of the semiconductor are influenced by the mixture of substances that would have dissolved in water. The electrical double layer that will exist if charged species were to be present in solution, then this state can affect the electron hole pair separation and also the adsorption properties of the dyes on the semiconductor surface. The rate of photocatalytic degradation of dyes depends on pH and the actual value of the pH at which the rate is maximum depending on the nature of the dye. In alkaline medium, hydroxyl radical ($\bullet\text{OH}$) an oxidant can be formed, thus increasing the rate of photodegradation of the dye [73].

Table 3. Data on the effect of pH on the photocatalytic degradation of dyes.

Dye type	Light source	Photo-catalyst	pH range	Optimum pH	Refs.
Orange G	UV	Sn/TiO ₂ /Ac	1.0-12.0	2.0	[88]
Methyl Orange	UV	Pt/TiO ₂	2.5-11.0	2.5	[89]
Orange G	Visible	N-TiO ₂	1.5-6.5	2.0	[90]
Acid Red B	UV	Ce-TiO ₂	1.5-7.0	1.5	[91]
Bromo-cresol purple	UV	TiO ₂	4.5 & 8.0	4.5	[64]
Methyl Red	visible	3% Ag+1.5% Ni-TiO ₂	3-10	4	[154]
Methylene Blue	UV	TiO ₂	3-9	Alkaline pH	[155]
Congo Red	Visible	Chitosan/CdS	6-12	6	[156]
Orange H	Solar	Zn-TiO ₂	3.0-10.0	3	[74]
Malachite Green	Sun light	Ni/MgFe ₂ O ₄	2.0-10.0	4	[157]
Indigo Carmine	UV	TiO ₂	4.0-11.0	4	[158]
Textile dye	UV	TiO ₂	3.0-7.0	5	[159]
Solophenyl Red 3B1	UV	TiO ₂	2.0-14.0	7	[160]
Methyl Orange	UV	Fe ² /C/S doped TiO ₂	2.0-12.0	Acidic Medium	[161]
Acid Orange 7	UV	WO ₃	3.0-9.0	3	[162]
Basic Yellow 28	UV	TiO ₂	3.0-9.0	5	[163]
Methylene Blue	UV	TiO ₂ ZnO	1.0-6.0	2	[164]
Reactive Blue 4	UV	Anatase TiO ₂	3.0-13.0	3-7	[165]
Methylene Blue	UV	ZnO	2.0-11.0	7(minimum)	[166]
Congo Red	UV	ZnO	5.0-10.0	8	[167]
Reactive Red 2	UV	TiO ₂	4.0-12.0	4-6	[168]
Procion Yellow	UV	TiO ₂	2.0-10.0	7.8	[169]
Acid Orange	UV	WO ₃ -TiO ₂	1.0-9.0	3.0	[7]
Acid Yellow	UV	TiO ₂	---	3.0	[6]
Amido Black 10B	UV	TiO ₂	----	9.0	[6]
Methyl Orange	UV	TiO ₂	2.0-10.0	8.0	[24]
Rhodamine B	UV	ZnO	2.0-12.0	12.0	[23]
Methyl Orange	Visible	Mg doped TiO ₂	3.0-8.0	--	[20]
Acid Orange 10	UV	TiO ₂	1.0-11.0	3.0	[170]
Methyl orange, Rhodamine B	UV	ZnO	2.0-10.0	Basic medium	[137]

In Table 3, some of the specific results on pH influence on the photodegradation of various dyes are given. Therefore, it is important to study the chemical nature of the dyes to be degraded and determine the correct pH to degrade them photo-catalytically. pH variation coupled with the temperature of calcination can result in the phase changes in the semiconductor, and hence the activity also may vary with pH. It is seen from the values given in Table 3 that at acidic pH, TiO_2 degrades the dyes effectively, that means that H^+ ions are mostly adsorbed on the semiconductor and is favourable for dye degradation. This also implies that certain forms of the dye alone are preferentially photo-degraded on TiO_2 surfaces.

The effect of pH on the photodegradation of dyes has to be rationalised based on the reaction mechanism. There are three possible ways dyes can degrade as a function of pH namely (i) direct attack by hydroxyl radical (ii) direct involvement of the positive hole in the oxidation reaction (iii) direct reduction by the participation of the electron excited to the conduction band.

A still more extensive compilation on the optimum pH on the photocatalytic degradation of dyes is already reported as Table.2 in the publication [83].

2.3. The Issues on Hand

Most of the published literature covers as variables, the light source, its intensity, pH of the medium, the amount of the catalyst employed, the initial concentration of the dye taken for study, the irradiation time and the other species like oxygen present in the reaction medium. Almost all the publications have been following these variables invariably. It is recognised that the study of these variables is important for assessing the utility of this method for pollutant removal (textile dye industry) from the waste water stream. The purpose of this presentation is to examine on what other aspects of these parameters can be intrinsically examined.

2.3.1. Kinetics of Photodegradation of Dyes

The kinetics of photocatalytic degradation of organic pollutants and dyes by semiconductors have been most often

treated as first order kinetics. This is most common in literature, and as an example one of the recent references [15, 54,55 and many other references] is provided. In Table.4. The literature data where the kinetics of degradation of dye has been treated according to pseudo first order kinetics are summarised.

The purpose is to analyse some of the consequences of treating the kinetic data on the removal of pollutants and other organic species especially under photo-catalytic conditions generally under First order kinetic equation. The first order kinetic equation employed in such circumstances [179] can be written as $-\ln(C/C_0) = kt$; where C is the concentration at any time t seconds, and C_0 is the initial concentration of the dye, and k is the value of the rate constant, this rate constant may be a lumped parameter including the value of the intrinsic rate constant, adsorption equilibrium constant and so on. Typical kinetic data analysed according to first order kinetic equation of the photo catalytic decomposition of Rhodamine B from ref 15 is given as an example.

The main conclusion of this study is that the inherent rate consists of the photo-catalytic and also photo-Induced self-degradation of the dye follows first order kinetics. If this argument were to be accepted then the treatment of kinetic data according to first order is only grossly approximate. Moreover, the apparent rate constant in the equation is a lumped parameter consisting of mostly the value of the intrinsic rate constant, the rates of other parallel reactions that would have taken place on the surface of the catalyst and many other accompanying non-elucidated rates of degradation. Possibly, the value of the apparent rate constant cannot be taken as a measure of the activity of the catalyst for comparison since the process taking place on the two or more catalysts are not identical or not even similar. This will have serious misconceptions for comparison purposes. In the example given, the authors report the apparent rate constant on the most active catalyst as 23.9 min^{-1} while the value of the apparent rate constant for the degradation of chlorophenol (where the photon induced degradation is assumed to be nearly negligible) is 3.47 min^{-1} which can be assumed in this case as the value of the intrinsic rate constant. May be cau-

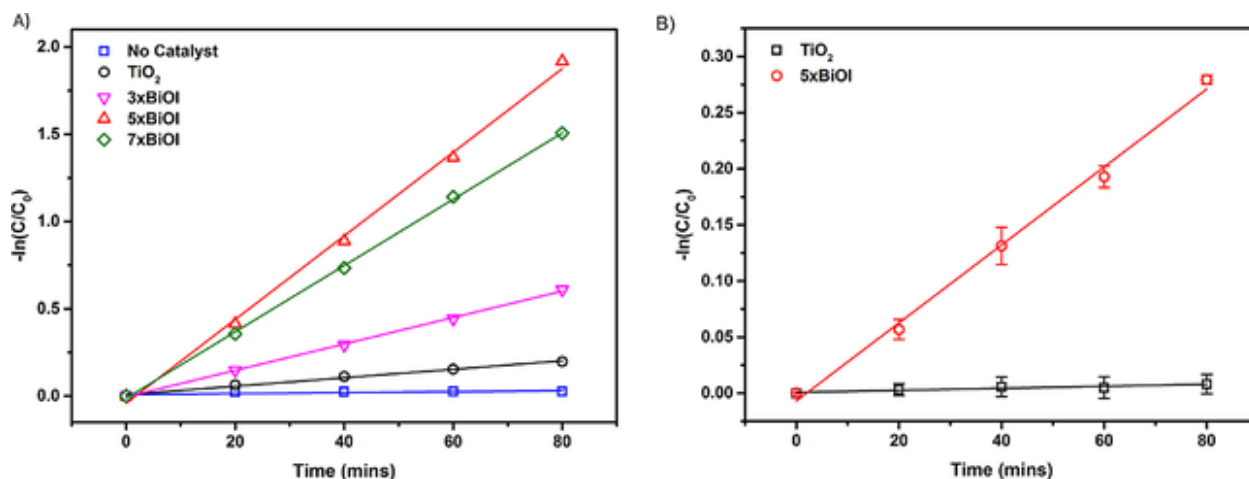


Fig. (6). Photocatalytic degradation kinetics of Rhodamine B on various photo-catalysts treated according to first order kinetic equation [data reproduced from ref.15. There is any number of this kind of analysis reported in the literature on the photocatalytic decomposition of dyes, and they are referred to in this article at other places] as an example.

Table.4. Values of first order rate constants for the photocatalytic degradation of dyes on semiconductor surfaces.

Semiconductor	Dye Degraded	Value for First Order Rate Constant	Refs.
ZnO	Methyl orange (MO)	0.1578 min ⁻¹	[171]
ZnO	Congo Red (CR)	0.1119 min ⁻¹	[171]
ZnO	Direct Black DB38)	0.0803 min ⁻¹	[171]
ZnO	Crystal Violet (CV)	0.079 min ⁻¹	[19]
ZnO	Methyl Red (MR)	0.014 min ⁻¹	[19]
ZnO	Basic Blue (BB)	0.1 min ⁻¹	[19]
TiO ₂	Crystal Violet (CV)	0.026 min ⁻¹	[19]
TiO ₂	Methyl Red (MR)	0.008 min ⁻¹	[19]
TiO ₂	Basic Blue (BB)	0.045 min ⁻¹	[19]
SnO ₂	Crystal Violet (CV)	0.010 min ⁻¹	[19]
SnO ₂	Methyl Red (MR)	0.004 min ⁻¹	[19]
SnO ₂	Basic Blue (BB)	0.017 min ⁻¹	[19]
Degussa P-25	Crystal Violet (CV)	0.060 min ⁻¹	[19]
Degussa P-25	Methyl Red (MR)	0.012 min ⁻¹	[19]
Degussa P-25	Basic Blue (BB)	0.017 min ⁻¹	[167]
ZnO Nano flowers	Methyl Orange (MO)	0.05485 min ⁻¹	[22]
ZnO Nano flowers	Congo Red	0.04611 min ⁻¹	[22]
ZnO Nano flowers	Chicago Sky blue	0.003182 min ⁻¹	[22]
ZnO Nano flowers	Eosin B	0.002884 min ⁻¹	[22]
3 nm GdCoO ₃	Rhodamine B	0.065 min ⁻¹	[172]
3 nm GdCoO ₃	Rhodamine Blur	0.078 min ⁻¹	[172]
3 nm GdCoO ₃	Orange G	0.053 min ⁻¹	[172]
3 nm GdCoO ₃	Remazol Brilliant Blue	0.019 min ⁻¹	[172]
TiO ₂ (450)	Methylene Blue	0.0102 min ⁻¹	[2]
TiO ₂ (450)	Congo Red	0.0085 min ⁻¹	[2]
TiO ₂	Reactive Red	0.0325 min ⁻¹	[168]
ZnO	Congo Red	0.0586 g.mg ⁻¹ .min ⁻¹	[167]
TiO ₂	Lissamine Green B	0.0165 min ⁻¹	[173]
ZnO	Methylene Blue	0.0135 min ⁻¹	[166]
WO ₃	Acid Orange 7	0.0225 min ⁻¹	[174]
Fe ³⁺ /C/S/TiO ₂	Methyl Orange	0.01628 min ⁻¹	[175]
Fe ³⁺ /C/S/TiO ₂	Congo Red	0.01533 min ⁻¹	[176]

TiO ₂	Carmin dye	0.1456 min ⁻¹	[177]
Chitosan/CdS	Congo Red	0.01108 min ⁻¹	[161]
ZnO	Methyl Orange	0.00029 sec ⁻¹	[178]
ZnO	Rhodamine 6G	0.00027sec ⁻¹	[178]
TiO ₂	Acid Orange 10	0.0326 min ⁻¹	[82]
TiO ₂	Acid Orange 12	0.0269 min ⁻¹	[82]
TiO ₂	Acid Orange 08	0.0235 min ⁻¹	[82]
TiO ₂	Amido-black- 10B	0.02083 min ⁻¹	[84]
TiO ₂	Methyl Red	0.0019 min ⁻¹	[85]
Ag1.5Ni0.75/TiO ₂	Methyl Red	0.0077 min ⁻¹	[85]
Ag1.5Ni1.5/TiO ₂	Methyl Red	0.0085 min ⁻¹	[85]
Ag1.5Ni3.0/TiO ₂	Methyl Red	0.009 min ⁻¹	[85]
Ag3.0Ni1.5/TiO ₂	Methyl Red	0.0111 min ⁻¹	[85]
TiO ₂ P25	Organic dye	0.003 to 0143 Min ⁻¹	[90]
ZnS	Rose Bengal	~4.51 X10 ⁻⁵ sec ⁻¹	[91]
CeCrO ₃	Fast Green	~4.41 X 10 ⁻⁴ sec ⁻¹	[92]
TiO ₂	RG 19	~4.69 h ⁻¹	[93]
TiO ₂	AO7	~2.07 h ⁻¹	[93]
TiO ₂ 0.02% Cu-doped TiO ₂ 0.04% Cu-doped TiO ₂ 0.06% Cu-doped TiO ₂ 0.08% Cu-doped TiO ₂ 0.1 % Cu-doped TiO ₂	Reactive Blue	0.0268 min ⁻¹ 0.0347 min ⁻¹ 0.03689 min ⁻¹ 0.0310 min ⁻¹ 0.0288 min ⁻¹ 0.0239 min ⁻¹	[94]
Au ³⁺ -doped SiO ₂ Ag ⁺ -doped SiO ₂ Ag-deposited SiO ₂ Au NP-&Ag SiO ₂ Au NP-SiO ₂ NPs SiO ₂ 2 NPs.	Methyl Red	0.370 min ⁻¹ 0.050 min ⁻¹ 0.046 min ⁻¹ 0.037 min ⁻¹ 0.032 min ⁻¹ 0.020 min ⁻¹	[86]

tion has to be exercised while comparing two or more catalytic systems on the basis of the rate constant values of the kinetic data treated as first order since on all catalyst systems the reaction may not follow the same kinetics though the treatment according to first order kinetics may apparently satisfy the first order kinetics. The statements given may apply to all general reactions which can involve multiple steps like preceding or succeeding surface reactions which are more often treated with first order kinetics. However, it is not our intention to make a general treatment.

Dye degradation can have many preconditions, one of them is the adsorption of the dye on the catalyst surface and this equilibrium constant should be reflected in the value of the rate constant evaluated from the data. The values of the equilibrium constants of adsorption on various catalyst surfaces can give the same or different order of reactivity of adsorbents, and this has to be considered while choosing the material for wastewater treatment.

The adsorption of dyes can follow any one of the known isotherms, like Langmuir or Freundlich isotherms. The rate

of decomposition of the dye depends on the adsorption equilibrium that exists on the surface and the adsorbed concentration (θ : surface coverage). Also, surface characteristics of the solid surface and the nature of the dye are the factors to be considered for adsorption. The pH of the solution also can change the surface charges and hence the adsorption of the dyes. Most of the dyes are considered as cationic or anionic, and the surface acidic or basic property will be responsible for these charged dyes. This is one aspect that has not been taken into account while treating the kinetics of photocatalytic degradation of dyes by semiconductors.

In most of the studies reported, details of adsorption of the dye and the consequences of this adsorption process on the kinetics of degradation have not been linked. This linking will be necessary for adoption of this technology for the treatment of waste water.

2.3.2. The Catalyst Loading

Another observation invariably recorded in literature is that the rate increases with catalyst loading till certain weight and above this the rate of degradation of the dye decreases with increase in weight. This is not an unusual result because the exposed surface area of the catalyst will not be directly proportional to the amount of catalyst loaded in solution phase reactions. Since dye degradation is proportional to the amount that is adsorbed on the surface of the solid, there can be a saturation point beyond which the solid amount may not have a direct relationship to the degradation extent. In most of the studies reported the maximum amount of the solid loaded for maximum activity is 3-4 g per litre [3] of the dye solution. This weight of the solid probably indicates the saturation limit of adsorption of the dye and possibly limits the concentration of the dye solution that can be employed for degradation, and thus the industries polluting waterways must restrict their pollution limits to this level. This may be a mark for pollution control authorities to note and it must restrict pollution to this level.

2.3.3. Effect of Light Intensity and Wavelength of Irradiation

Use of solar radiation through less expensive and less hazardous, reproducible results are not assured. The effect of light intensity on the photocatalytic performance has been identified in three ranges, In the case of low intensity, the rate increases with intensity while in the intermediate intensity range the rate increases as the square root of the intensity and high intensity the rate is independent of the light intensity. On the whole, the rate of degradation is better when the light is from UV-Visible sources rather than solar radiation. Secondly, the electron hole pair formation and their recombination may be different when the UV-Visible source is employed. The wavelength of the light source can also affect the electron hole formation rate and their recombination.

2.3.4. The Mechanism of Dye Degradation

The degradation studies simply measure optical absorption (decolourization) or COD and thus not all information is available on the fragments and degradation products except that carbon dioxide and water are formed ultimately from the degradation process. This situation can arise because of anxiety on the removal of pollutant from water, and it is

assumed that the degradation products are not harmful to living beings. The degradation scheme of a typical acid orange 7 is shown in Fig. 7. In this case, the final product appears as carbon dioxide and water, but there are some intermediates which may participate in the reaction scheme of degradation of the dye and thus affect the overall kinetics of degradation. Among the various dyes studied, the extensively studied system is Acid Dye7 (AO 7) by some workers [89, 71-75]. A variety of degradation products has been identified, and the main ones are benzene sulphonic acid, sulphonic acid, 1,4 naphthoquinone and phthalic acid as by-products. Also, various other products have also been identified in the degradation of this dye; these include 2-naphthol, 2-hydroxy-1,4 naphthoquinone and small amounts of phthalimide, aliphatic acids like fumaric and succinic maleic and malonic acids together with other lower molecular weight products. Taking these observations, the reaction sequence for the photooxidation of AO 7 para isomer (Acid Orange 20) was proposed and is shown in Fig. 7. This scheme is applicable only for the catalytic (TiO₂) photo-degradation conditions. The main oxidising species are superoxide anion radical or hydroxyl radicals that can be formed by the participation of the exciton formed in the semiconductor photo-catalyst under irradiated conditions.

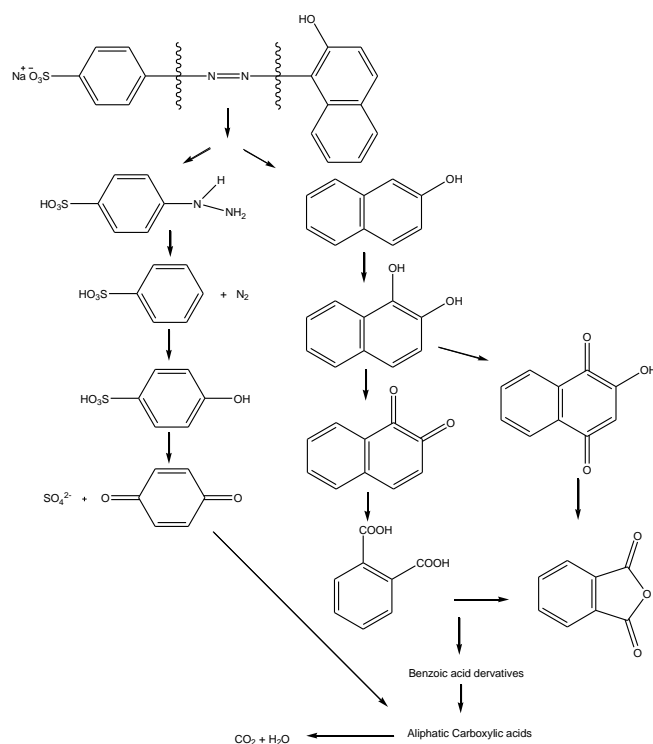


Fig. (7). Possible photo-catalytic pathway for the degradation of acid orange 7 [deduced from the results reported in the publications 71-75].

Most of the studies reported in literature either measure spectrophotometrically decolourization or COD. These are only gross measurements, and details of the degradation and the nature of degraded fragments have not been identified in most of the studies reported. Therefore, there is a need to study in detail the degradation products in the case of a number of dyes employed.

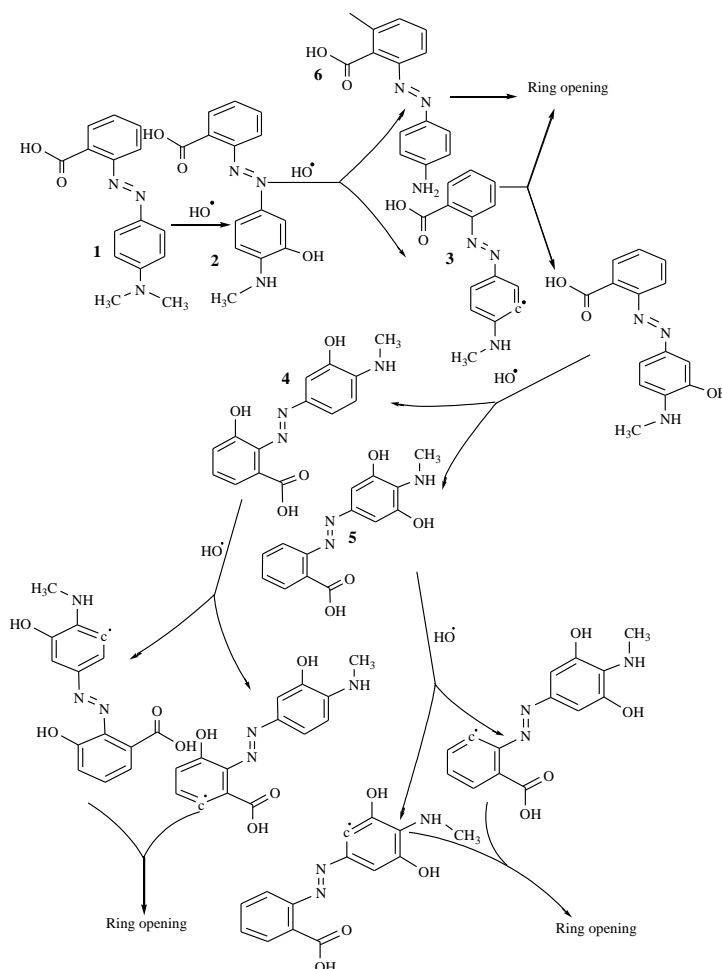


Fig. (8). Mechanism of the possible routes for the photocatalytic degradation of methyl Red [Reproduced from ref 86].

Methyl Red decomposition has been studied by Mahmoud *et al.* [86], and the main species responsible for this degradation is identified as superoxide anion radical ($O_2^{\cdot-}$), or hydroxide radical ($\cdot OH$) and the degradation route proposed by them is given in Fig. 8. The possible first step can be the attack of the $\cdot OH$ radical on methyl Red which leads to the formation of a dehydrogenated radical. This intermediate can either undergo ring opening combining with $\cdot OH$, forming hydroxyl product. The intermediate might also be decomposed to form a new low molecular weight by-product. Further attacked by $\cdot OH$ to form bi-hydroxyl products. The same procedure could take place until complete ring opening, and complete mineralisation occurs. The proposed sequence of steps is schematically shown in Fig. 8.

One of the conventional azo dye is acid orange 7 whose proposed decomposition route as elucidated in literature is shown in Fig. 9. One can see that the bond breaking either by dealkylation or hydroxylation of the dye derivatives and thus finally lead to mineralisation of the dye.

2.3.5. Reactive Oxygen Species

In the earlier parts, intrinsic reactive oxygen species (ROS) namely hydroxyl radical ($\cdot OH$), hydrogen peroxide (H_2O_2), superoxide anion radical ($\cdot O_2^-$) and singlet oxygen (1O_2), have been identified as the reactive oxygen species responsible for the dye degradation. The adsorbed hydroxyl radical ($\cdot OH$)

could be regarded as trapping holes by hydroxyl species involved in the rapid adsorption-desorption equilibrium at the semiconductor-solution interface. The trapped holes must be the dominant oxidation species because the equilibrium shifts to the adsorption side whereas $\cdot OH$ in solution would exert the reactivity towards the non-adsorbed reactants. The most probable routes for generating these species on semiconductor surfaces have been discussed elsewhere [89]. These authors have given the pictorial representation of the processes, and the same is reproduced in Fig. 10.

Even though the nature of the reactive active species has been identified, their exact origin and participation in photocatalytic degradation of dyes on various semiconductors have not yet been fully elucidated. It is possible that further research in this area will throw light on this important area of research.

CONCLUDING REMARKS

It is true that photo-catalytic degradation of dyes has been studied throughout the world. This activity is intensified in the recent years [180-234]. These studies reported in literature deal with the common variables as stated above. In addition to summarising the results of these studies, this presentation focuses on certain aspects that can also be explored. These include:

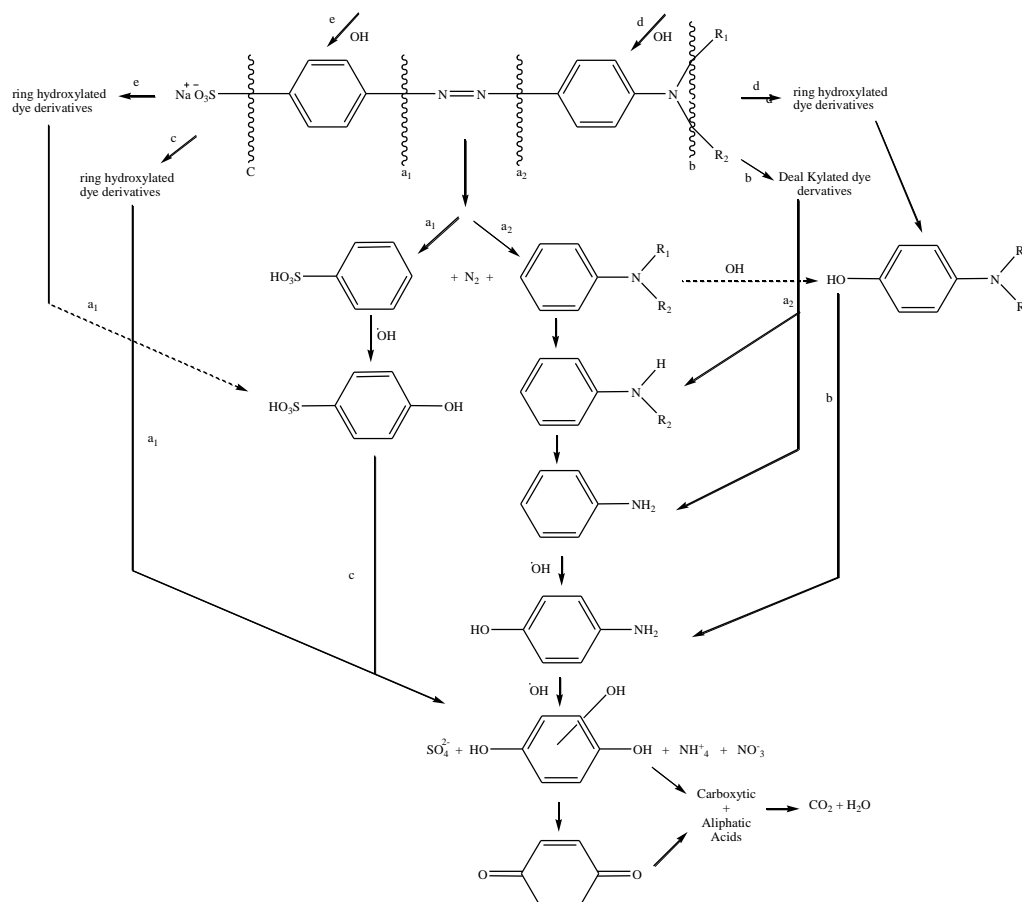


Fig. (9). Routes of the degradation of a typical azo dye namely acid orange 7 [fig reproduced from ref.54].



Fig. (10). Pictorial representation of generation of Reactive Oxygen Species on photo-catalytic semiconductor surfaces.

1. The inadequacy of treating the kinetic data in terms of pseudo-first order rate. It is necessary to take into account the prior adsorption and characteristics of this process.
2. The optimum amount of the catalyst is a natural consequence, and in this, the identification of the active sites and their number density on semiconductor surfaces can be evaluated.
3. The degradation of dye is dependent on the nature of the oxidising agent used, and the choice of the oxidising agent must be considered in terms of the redox potential of the species.

4. The pH of the medium decides which form of the dye is adsorbed on the surface and how it is degraded. The variation with pH has to be visualised in terms of the surface characteristics of the semiconductor as a function of pH.

It is concluded that the variables studied in the photocatalytic degradation of dyes are appropriate ones but the studies can extend their scope, and the interpretation can cover more global significance.

CONSENT FOR PUBLICATION

Not applicable.

CONFLICT OF INTEREST

The authors declare no conflict of interest, financial or otherwise.

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